# Tucannon River Spring Chinook Salmon Hatchery Evaluation Program 

## 2002 Annual Report

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## Abstract

Lyons Ferry Hatchery (LFH) and Tucannon Fish Hatchery (TFH) were built/modified under the Lower Snake River Fish and Wildlife Compensation Plan. One objective was to compensate for the estimated annual loss of 1,152 -spring chinook (Tucannon River stock) caused by hydroelectric projects on the Snake River. The standard supplementation production goal is 132,000 fish for release as yearlings at $30 \mathrm{~g} /$ fish or 15 fish per pound ( fpp ). The captive brood production goal is 150,000 yearlings at 30 g /fish. This report summarizes activities of the Washington Department of Fish and Wildlife Lower Snake River Hatchery Evaluation Program for Tucannon River spring chinook for the period April 2002 to April 2003.

Seven hundred seventy-eight fish were captured in the TFH trap in 2002 (166 natural adults, 2 natural jacks, 589 hatchery adults, and 21 hatchery jacks); 107 were collected and hauled to LFH for broodstock and the remaining fish were passed upstream.

During 2002, two salmon that were collected for broodstock died. Prespawning mortality has been low since broodstock began being held at LFH in 1992, and is generally less than $10 \%$ each year.

Spawning of supplementation fish in 2002 at LFH occurred between August 27 and September 17, with peak eggtake on September 3. A total of 169,364 eggs were collected from 22 wild and 25 hatchery-origin fish. Egg mortality to eye-up was $3.6 \%$ ( 6,047 eggs), with an additional loss of $11,786(7.2 \%)$ sac-fry. Total fry ponded for production in the rearing ponds was 151,531 .

A total of 121 captive brood females were spawned from August 27 to October 2, 2002 producing 176,544 eggs. Egg mortality to eye-up was $68 \%$ leaving 55,711 live eggs. An additional 5,249 dead eggs/fry (9.4\%) were picked at ponding leaving 50,462 fish for rearing.

One spring chinook salmon that was radio tagged at Bonneville Dam entered the Tucannon River in 2002. This fish entered the adult trap on June 12 and was passed upstream by hatchery personnel. Efforts to locate this fish after it was passed upstream were unsuccessful. The radio tag either quit working or the fish/transmitter left the area.

On August 20, 97 excess captive broodstock were released at Panjab Bridge. All released fish were Monel jaw tagged and radio transmitters were inserted into ten of the largest fish. Only two of the radio tagged females spawned and released captive brood were observed being dominated by hatchery and wild fish in the river.

WDFW staff conducted spawning ground surveys in the Tucannon River between August 22 and October 7, 2002. One hundred ninety-seven redds and 140 carcasses were found above the adult trap and 102 redds and 60 carcasses were found below the trap. Based on redd counts,
broodstock collection, and in-river pre-spawning mortalities, the estimated escapement for 2002 was 1,005 fish ( 341 wild adults, 9 wild jacks and 644 hatchery-origin adults, 11 hatchery jacks).

Length and weight samples were collected three times during the rearing cycle for 2001 BY juveniles at TFH and Curl Lake Acclimation Pond. All 2001 BY juveniles were marked in October at LFH, transported to TFH, and transported again in February to Curl Lake for acclimation and volitional release during March and April.

Snorkel surveys were conducted during the summer of 2002 to determine the population of subyearling and yearling spring chinook in the Tucannon River. We estimated 63,412 subyearlings (BY 2001) and 703 yearlings (BY 2000) were present in the river. Evaluation staff also operated a downstream migrant trap. During the 2001/2002 emigration, we estimated that 20,049 (BY 2000) wild spring chinook smolts emigrated from the Tucannon River.

Monitoring survival rate differences between natural and hatchery-reared salmon continues. Smolt-to-adult return rates (SAR) for natural salmon consistently average about three times higher than for hatchery salmon. However, hatchery salmon survive about four times greater than natural salmon from parent to adult progeny. Due to the low SAR for hatchery fish, the mitigation goal of 1,152 salmon of Tucannon River stock was not achieved.

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## Program Objectives

Congress authorized implementation of the Lower Snake River Fish and Wildlife Compensation Plan (USACE 1975). As a result, Lyons Ferry Hatchery (LFH) was constructed and Tucannon Fish Hatchery (TFH) was modified. One objective of these hatcheries is to compensate for the estimated annual loss of 1,152 Tucannon River spring chinook salmon adults caused by hydroelectric projects on the Snake River. In 1984, Washington Department of Fish and Wildlife (WDFW) began to evaluate the success of these two hatcheries in meeting the mitigation goal, and identifying factors that would improve performance of the hatchery fish. The WDFW also initiated the Tucannon River Spring Chinook Captive Broodstock Program in 1997 that is currently funded by the Bonneville Power Administration (BPA). The project goal is to rear captive salmon selected from the supplementation program (1997-2001 BYs) to adults, rear their progeny, and release approximately 150,000 smolts ( $30 \mathrm{~g} /$ fish) annually into the Tucannon River between 2003-2007. These smolt releases, in combination with the current hatchery supplementation program (goal $=132,000$ smolts) and wild production, are expected to produce 600-700 returning adult spring chinook to the Tucannon River each year from 20052010. This report summarizes work performed by the WDFW Spring Chinook Evaluation Program from April 2002 through April 2003.

## Facility Descriptions

Lyons Ferry Hatchery is located on the Snake River (rkm 90) at its confluence with the Palouse River (Figure 1). It is used for adult broodstock holding and spawning, and early life incubation and rearing. All juvenile fish are marked and returned to TFH for final rearing and acclimation. Tucannon Fish Hatchery, located at rkm 59 on the Tucannon River, has an adult collection trap on site (Figure 1). Juveniles rear at TFH through winter. In February, the fish are transported to Curl Lake Acclimation Pond (AP) and volitionally released. The annual supplementation production goal is 132,000 fish for release as yearlings at 30 g /fish or 15 fish per pound (fpp). The captive brood production goal is 150,000 yearling smolts at $30 \mathrm{~g} /$ fish.

## Tucannon River Watershed Characteristics

The Tucannon River empties into the Snake River between Little Goose and Lower Monumental dams approximately 622 rkm from the mouth of the Columbia River (Figure 1). Stream elevation rises from 150 m at the mouth to $1,640 \mathrm{~m}$ at the headwaters (Bugert et al. 1990). Total watershed area is approximately $1,295 \mathrm{~km}^{2}$. Local habitat problems related to logging, road building, recreation, and agriculture/livestock grazing have limited the production potential of
spring chinook in the Tucannon River. Land use in the Tucannon watershed is approximately $36 \%$ grazed rangeland, $33 \%$ dry cropland, $23 \%$ forest, $6 \%$ WDFW, and $2 \%$ other use (Tucannon Subbasin Summary 2001). Five unique strata have been distinguished by predominant land use, habitat, and landmarks (Table 1).


Figure 1. Location of the Tucannon River, Lyons Ferry, and Tucannon hatcheries within the Snake River Basin.

Table 1. Description of five strata within the Tucannon River.

| Strata | Land Ownership/Usage | Spring Chinook Habitat | River Kilometer ${ }^{\mathbf{a}}$ |
| :---: | :---: | :---: | :---: |
| Lower | Private/Agriculture \& Ranching | Not-Usable (temperature limited) | $0.0-20.1$ |
| Marengo | Private/Agriculture \& Ranching | Marginal (temperature limited) | $20.1-39.9$ |
| Hartsock | Private/Agriculture \& Ranching | Fair to Good | $39.9-55.5$ |
| HMA | State \& Forest Service/Recreational | Good/Excellent | $55.5-74.5$ |
| Wilderness | Forest Service/Recreational | Excellent | $74.5-86.3$ |

${ }^{\text {a }}$ Rkm descriptions: 0.0-mouth at the Snake River; 20.1-Territorial Rd.; 39.9-Marengo Br.; 55.5-HMA Boundary Fence; 74.5-Panjab Br.; 86.3-Rucherts Camp.

Program staff deployed 19 continuous recording thermographs throughout the Tucannon River to monitor daily minimum and maximum water temperatures (temperatures are recorded every 1 to
1.2 hours) from May through October. Data from each of these water temperature recorders are kept on an electronic file in our Dayton office. During 2002, maximum temperatures near the mouth (rkm 3) of the Tucannon River reached 26.7 C ( 80 F ) on 3 different days. Maximum temperatures where spring chinook juveniles were rearing during the hottest part of the summer ranged from 15.9 C ( 60.7 F ) in the upper HMA stratum (rkm 74.5) to 23.1 C (73.6 F) in the lower Hartsock stratum (rkm 43.3)(Figure 2).

The upper lethal temperature for chinook fry is $25.1 \mathrm{C}(77.2 \mathrm{~F})$ while the preferred temperature range is $12-14 \mathrm{C}(53.6-57.2 \mathrm{~F}$ ) (Scott and Crossman 1973). The optimum range of temperature in freshwater, which controls the rate of growth and survival of young, is 13-17 C (55.4-62.6 F) (Becker 1983). Theurer et al. (1985) estimated that spring chinook production in the Tucannon River would be zero for all stream reaches having maximum daily July water temperatures greater than $23.9 \mathrm{C}(75 \mathrm{~F}$ ) (or average mean temperature of $20 \mathrm{C}(68.0 \mathrm{~F})$ ). Based on the preferred and optimum temperature limits, fish returning to the upper watershed have the best chance for survival (Figure 2).

It is hoped that recent initiatives to improve habitat within the Tucannon Basin, such as the Tucannon River Model Watershed Program, will: 1) restore and maintain natural stream stability; 2) reduce water temperatures; 3) reduce upland erosion and sediment delivery rates; and 4) improve and re-establish riparian vegetation. Theurer et al. (1985) estimated that improving riparian cover and channel morphology in the Tucannon River mainstem would increase chinook-rearing capacity present in the early 1980s by a factor of 2.5 . Habitat restoration efforts should permit increased utilization of habitat by spring chinook salmon in the marginal sections of the middle reaches of the Tucannon River and increase fish survival.


Figure 2. Maximum temperature, average maximum temperature, and average minimum temperature recorded by thermographs at 19 selected sites along the Tucannon River, May-October, 2002.

## Adult Salmon Evaluation

## Broodstock Trapping

The annual collection goal for broodstock is 50 natural and 50 hatchery adults collected throughout the duration of the run. Additional jack salmon may be collected to contribute to the broodstock if necessary. Jack contribution to the broodstock can be no more than their percentage in the overall run. Returning hatchery salmon were identified by lack of the adipose fin.

The TFH adult trap began operation in February (for steelhead) with the first spring chinook captured May 4. The trap was operated through September. A total of 778 fish entered the trap ( 166 natural adults, 2 natural jacks, 589 hatchery adults, and 21 hatchery jacks), and 107 were collected and hauled to LFH for broodstock (Table 2, Appendix A). Fish not collected for broodstock were passed upstream. Adults collected for broodstock were injected with erythromycin and oxytetracycline ( $0.5 \mathrm{cc} / 4.5 \mathrm{~kg}$ ); jacks were given half dosages. Fish received formalin drip treatments during holding at 167 ppm every other day at LFH to control fungus.

Based on previous year returns, we anticipated catching unmarked Umatilla origin hatchery fish. We decided prior to broodstock trapping that scale samples would be collected from all unmarked fish for scale pattern analysis in the hope of identifying hatchery origin fish. Unmarked fish collected for broodstock were injected with a Passive Integrated Transponder (PIT) tag for individual identification. If scale analysis determined that a "wild" fish collected for broodstock was actually of hatchery origin, that fish would be identified by its PIT tag number and killed. Ten of the "wild" fish collected for broodstock were determined to be of hatchery origin and were killed to protect genetic integrity. These ten fish were ripe during the first spawn day, which tends to confirm our suspicions that they were from the Umatilla River as that stock spawns earlier than Tucannon origin fish. To ensure that we would reach our eggtake goal, four additional hatchery fish (visible AD clip) were collected at the adult trap during September.

Table 2. Numbers of spring chinook salmon captured, trap mortalities, fish collected for broodstock, or passed upstream to spawn naturally at the TFH trap from 1986-2002.

| Year | Captured at Trap |  | Trap Mortality |  | Broodstock Collected |  | Passed Upstream |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Natural | Hatchery | Natural | Hatchery | Natural | Hatchery | Natural | Hatchery |
| 1986 | 247 | 0 | 0 | 0 | 116 | 0 | 131 | 0 |
| 1987 | 209 | 0 | 0 | 0 | 101 | 0 | 108 | 0 |
| 1988 | 267 | 9 | 0 | 0 | 116 | 9 | 151 | 0 |
| 1989 | 156 | 102 | 0 | 0 | 67 | 102 | 89 | 0 |
| 1990 | 252 | 216 | 0 | 1 | 60 | 75 | 191 | 134 |
| 1991 | 109 | 202 | 0 | 0 | 41 | 89 | 68 | 105 |
| 1992 | 242 | 305 | 8 | 3 | 47 | 50 | 165 | 202 |
| 1993 | 191 | 257 | 0 | 0 | 50 | 47 | 130 | 167 |
| 1994 | 36 | 34 | 0 | 0 | 36 | 34 | 0 | 0 |
| 1995 | 10 | 33 | 0 | 0 | 10 | 33 | 0 | 0 |
| 1996 | 76 | 59 | 1 | 4 | 35 | 45 | 33 | 7 |
| 1997 | 99 | 160 | 0 | 0 | 43 | 54 | 47 | 76 |
| $1998{ }^{\text {a }}$ | 50 | 43 | 0 | 0 | 48 | 41 | 1 | 1 |
| $1999{ }^{\text {b }}$ | 1 | 139 | 0 | 1 | 1 | 135 | 0 | 0 |
| $2000{ }^{\text {c }}$ | 28 | 177 | 0 | 17 | 12 | 69 | 13 | 94 |
| 2001 | 405 | 276 | 0 | 0 | 52 | 54 | 353 | 222 |
| 2002 | 168 | 610 | 0 | 0 | 42 | 65 | 126 | 545 |

${ }^{a}$ Two males (one natural, one hatchery) captured were transported back downstream to spawn in the river.
${ }^{\mathrm{b}}$ Three hatchery males that were captured were transported back downstream to spawn in the river.
${ }^{\text {c }} 17$ stray LV and ADLV fish were killed at the trap.

## Broodstock Mortality

Two of the 107 salmon collected for broodstock died prior to spawning in 2002 (Table 3). Table 3 shows that prespawning mortality in 2002 was comparable to the mortality documented since broodstock began being held at LFH in 1992. Higher mortality was experienced when fish were held at TFH (1986-1991).

Table 3. Numbers of prespawning mortalities and percent of fish collected for broodstock at TFH and held at TFH (1985-1991) or LFH (1992-2002).

| Year | Natural |  |  |  | Hatchery |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Male | Female | Jack | Male | Female | Jack | \% of collected |  |
|  | 3 | 10 | 0 | 59.1 | - | - | - | - |
| 1986 | 15 | 10 | 0 | - | - | - |  |  |
| 1987 | 10 | 8 | 0 | 17.8 | - | - | - | - |
| 1988 | 7 | 22 | 0 | 25.0 | - | - | - | 100.0 |
| 1989 | 8 | 3 | 1 | 17.9 | 5 | 8 | 22 | 34.3 |
| 1990 | 12 | 6 | 0 | 30.0 | 14 | 22 | 3 | 52.0 |
| 1991 | 0 | 0 | 1 | 2.4 | 8 | 17 | 32 | 64.0 |
| 1992 | 0 | 4 | 0 | 8.2 | 2 | 0 | 0 | 4.0 |
| 1993 | 1 | 2 | 0 | 6.0 | 2 | 1 | 0 | 6.4 |
| 1994 | 1 | 0 | 0 | 2.8 | 0 | 0 | 0 | 0.0 |
| 1995 | 1 | 0 | 0 | 10.0 | 0 | 0 | 3 | 9.1 |
| 1996 | 0 | 2 | 0 | 5.7 | 2 | 1 | 0 | 6.7 |
| 1997 | 0 | 4 | 0 | 9.3 | 2 | 2 | 0 | 7.4 |
| 1998 | 1 | 2 | 0 | 6.3 | 0 | 0 | 0 | 0.0 |
| 1999 | 0 | 0 | 0 | 0.0 | 3 | 1 | 1 | 3.8 |
| 2000 | 0 | 0 | 0 | 0.0 | 1 | 2 | 0 | 3.7 |
| 2001 | 0 | 0 | 0 | 0.0 | 0 | 0 | 0 | 0.0 |
| 2002 | 0 | 0 | 0 | 0.0 | 1 | 1 | 0 | 3.1 |

## Broodstock Spawning

Spawning at LFH occurred once a week from August 27 to September 17, with peak eggtake on September 3. A total of 169,364 eggs were collected (Table 4). Eggs were initially disinfected and water hardened for one hour in iodophor ( 100 ppm ). Fungus on the incubating eggs was controlled with formalin applied every-other day at $1,667 \mathrm{ppm}$ for 15 minutes. Mortality to eyeup was $3.6 \%$ with an additional $7.2 \%(11,786)$ loss of sac-fry, which left 151,531 fish for production.

To prevent any stray fish from contributing to the population, all coded wire tags (CWT) were read prior to spawning. Two hatchery females were determined to be strays and were killed outright. Scales from unmarked fish were read prior to spawning to check for hatchery growth patterns. Carcasses were buried instead of being used for nutrient enhancement due to the detection of Infectious Hematopoietic Necrosis virus in the broodstock.

Table 4. Number of fish spawned, estimated egg collection, and egg mortality of Tucannon River spring chinook salmon at LFH in 2002.

| Spawn Date | Natural |  | Hatchery |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Male | Female | Eggs Taken | Male | Female |


| $8 / 27$ | $4^{\mathrm{a}}$ | 3 | 10,858 | 4 | $4^{\mathrm{b}}$ | 6,702 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| $9 / 03$ | $14^{\mathrm{a}}$ | 6 | 26,264 | 6 | 13 | 43,004 |
| $9 / 10$ | $4^{\mathrm{a}}$ | 8 | 31,973 | 8 | 4 | 13,887 |
| $9 / 17$ | $6^{\mathrm{a}}$ | 5 | 17,021 | 8 | 6 | 19,655 |
| Totals | $\mathbf{2 0}$ | $\mathbf{2 2}$ | $\mathbf{8 6 , 1 1 6}$ | $\mathbf{2 6}$ | $\mathbf{2 5}$ | $\mathbf{8 3 , 2 4 8}$ |
| Egg Mortality |  |  | 2,041 |  |  | 4,006 |

${ }^{a}$ Denotes live spawned fish.
${ }^{\mathrm{b}}$ Two of the four hatchery females were determined to be strays and their eggs were destroyed.

Eggs were also collected as part of the Tucannon River Captive Broodstock Program. Based on our projections, if all mature captive broodstock were spawned we would exceed our eggtake goal again as happened in 2001. To prevent having excess fish on hand, we decided to outplant excess mature captive broodstock in 2002. From the 1998 and 1999 brood years, we removed 21 and 76 fish respectively for outplanting in the Tucannon River (see Radio Tracking Section). All other mature captive brood fish were spawned from August 27 to October 2. A total of 121 captive brood females were spawned for a total eggtake of 176,544 . Loss to eye-up was $68 \%$ leaving only 55,711 live eggs. Reasons for the high mortality are unknown but are most likely related to poor egg viability. An additional 5,249 dead eggs/fry were picked at ponding leaving only 50,462 fish for rearing. This is well below the program release goal ( 150,000 smolts) due to the higher than expected egg mortality. The Tucannon River Captive Broodstock Program was funded through the BPA and results achieved to date are more thoroughly described in the annual Tucannon River Spring Chinook Captive Broodstock Report (Gallinat and Varney 2003).

## Radio Tracking

A spring chinook that was radio tagged (channel 11, code 92 , frequency 149.52 MHz ) by the University of Idaho at Bonneville Dam was tracked in the Tucannon River during 2002. This fish passed our smolt trap (rkm 2.7) on May 15. On May 29 the fish was detected just below the Tucannon Hatchery Adult Trap (rkm 59). It entered the adult trap on June 12 and was passed upstream by hatchery personnel. Efforts to locate this fish after it was passed upstream were unsuccessful. The radio tag either quit working or the fish/transmitter left the area.

On August 20, 97 (21 1998 BY and 761999 BY) excess Tucannon River captive brood adult spring chinook were released (Table 5) into the Tucannon River at Panjab Bridge (rkm 74.5). All released fish were Monel jaw tagged and radio transmitters were inserted into ten of the larger (presumably female) fish for tracking and monitoring in the wild. Radio tagged fish were monitored weekly through the end of September (Appendix B). Table 5 summarizes the tagging and recovery information from the radio tagged fish. Two of the radio tagged females spawned successfully within 2 km of the release site ( $9 / 165$ and $9 / 192$ ). Another female (9/167) that was attempting to spawn (actively digging a redd) died after releasing less than $10 \%$ of her eggs. Of the remaining seven fish: three tags were recovered hidden on the stream bank without a carcass and may have been illegally harvested; two fish were eaten by predators; one fish was a prespawn mortality unrecoverable in a debris jam; and one fish (9/203) was never located after release - the radio stopped transmitting or the fish and transmitter left the area.

Table 5. Radio tagging and recovery data for ten adult captive spring chinook salmon tagged on July 16 and released on August 20 at Panjab Bridge in the Tucannon River during 2002.

|  | Release Data |  |  |  |  | Recovery Data |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Channel/ <br> Code | Panjab Br. <br> Rkm | Sex | FL (cm) | Recovery Information | Date | Rkm | Likely <br> Poached | Spawned |
| $9 / 165$ | 74.5 | F | 58.0 | Recovered fish \& tag | $9 / 25$ | 72.9 | No | Yes |
| $9 / 167$ | 74.5 | F | 55.5 | Recovered fish \& tag | $9 / 13$ | 73.0 | No | No |
| $9 / 171$ | 74.5 | F | 56.5 | Recovered fish \& tag | $9 / 23$ | 73.4 | No | No |
| $9 / 179$ | 74.5 | F | 55.5 | Tag found on bank | $9 / 20$ | 77.7 | Yes | No |
| $9 / 183$ | 74.5 | F | 52.0 | Tag found on bank | $9 / 20$ | 74.5 | Yes | No |
| $9 / 184$ | 74.5 | F | 51.0 | Carcass in log jam | -- | 68.7 | No | No |
| $9 / 192$ | 74.5 | F | 50.0 | Recovered fish \& tag | $9 / 27$ | 73.6 | No | Yes |
| $9 / 193$ | 74.5 | F | 51.0 | Tag in animal den | -- | 73.5 | No | No |
| $9 / 203$ | 74.5 | F | 49.0 | Lost contact | -- | -- | $? ?$ | $? ?$ |
| $9 / 205$ | 74.5 | F | 47.0 | Tag found on bank | $9 / 13$ | 76.6 | Yes | No |

Outplanted adults differed from wild and hatchery-origin fish in the river in morphology and coloration. Captive brood males lacked a prominent kype and captive fish were more goldenyellow in color. During redd surveys, released captive brood adults were observed being chased by more dominant male and female wild and hatchery-origin fish in the river.

In studies by Berejikian et al. (1997), wild coho females produced more nests than captive brood females. They also found that captive brood coho males were dominated by wild males and were also attacked more often by females. Fleming and Gross (1993) found coho hatchery females were delayed in spawning, retained more eggs, spawned in less desirable areas, and were less successful in guarding nest sites.

Losses to predation may be higher for fish released from a hatchery environment due to inappropriate avoidance and foraging behaviors and their inability to accurately assess predation risks, secondary stress effects, and a general unfamiliarity with their new surroundings (Steward and Bjornn 1990).

Due to the low frequency of natural spawning by released fish, high mortality due to predation and probable illegal harvest, and high egg mortality in the hatchery during 2002, priority will be to release excess progeny as parr to stay within smolt release goals rather than the release of excess captive broodstock as adults.

## Natural Spawning

Spawning ground surveys were conducted on the Tucannon River weekly from August 22 to October 7, 2002 to count redds and determine the temporal and spatial distribution of spawners. Two hundred ninety-nine redds were counted and 79 natural and 121 hatchery origin carcasses were recovered (Table 6). One hundred ninety-seven redds ( $66 \%$ of total) and 140 carcasses ( $70 \%$ of total) were found above the adult trap. U.S. Forest Service personnel reported seeing 5 chinook redds with no fish associated with them near Ruchert's Camp (rkm 86.3) (Bill Dowdy, U.S. Forest Service, personal communication). These may have been bull trout redds due to their location and lack of carcasses and were not included in the totals.

| Stratum | Rkm ${ }^{\text {a }}$ | Number of redds | Carcasses Recovered |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Natural | Hatchery |
| Wilderness | 78-84 | 3 | 0 | 1 |
| HMA | 75-78 | 10 | 0 | 3 |
|  | 73-75 | 46 | 10 | 19 |
|  | 68-73 | 61 | 7 | 40 |
|  | 66-68 | 22 | 2 | 8 |
|  | 62-66 | 43 | 12 | 14 |
|  | 59-62 | 12 | 10 | 14 |
| Hartsock | 56-59 | 43 | 24 | 13 |
|  | 52-56 | 24 | 9 | 7 |
|  | 47-52 | 12 | 3 | 0 |
|  | 43-47 | 4 | 1 | 1 |
| Marengo | 40-43 | 6 | 1 | 0 |
|  | 34-40 | 11 | 0 | 1 |
|  | 28-34 | 2 | 0 | 0 |
| Totals | 28-84 | 299 | 79 | 121 |
| ${ }^{\text {a }}$ Rkm descriptions: 84-Sheep Cr.; 78-Lady Bug Flat CG; 75-Panjab Br.; 73-Cow Camp Bridge; 68-Tucannon CG; 66-Curl Lake; 62-Beaver/Watson Lakes Br.; 59-Tucannon Hatchery Intake/Adult Trap; 56-HMA Boundary Fence; 52-Br. 14; 47-Br. 12; 43-Br. 10; 40-Marengo Br.; 34-King Grade Br.; 28-Enrich Br. |  |  |  |  |

## Historical Trends

Two general trends were evident from the program's inception in 1985 through 1999:

1) The proportion of the total number of redds occurring below the trap increased; and
2) The density of redds (redds $/ \mathrm{km}$ ) decreased in the Tucannon River.

In part, this resulted from a greater emphasis on broodstock collection to keep the spring chinook population above extinction. However, increases in the SAR rates beginning with the 1995 brood have subsequently resulted in increased spawning above the trap and higher redd densities. The number of redds in 2002 increased $225 \%$ from 2000 levels and were the most recorded since surveys began in 1985 (Table 7).

Table 7. Number of spring chinook salmon redds and redds/km (in parenthesis) by stratum and year, and the

|  | Strata |  |  |  | Total Redds | TFH Adult Trap |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Wilderness | HMA | Hartsock | Marengo |  | Above | \% | Below | \% |
| 1985 | 97 (8.2) | 122 (6.2) | - | - | 219 | - | - | - | - |
| 1986 | 53 (4.5) | 117 (6.2) | 29 (1.9) | 0 (0.0) | 200 | 163 | 81.5 | 37 | 18.5 |
| 1987 | 15 (1.3) | 140 (7.4) | 30 (1.9) | - | 185 | 149 | 80.5 | 36 | 19.5 |
| 1988 | 18 (1.5) | 79 (4.2) | 20 (1.3) | - | 117 | 90 | 76.9 | 27 | 23.1 |
| 1989 | 29 (2.5) | 54 (2.8) | 23 (1.5) | - | 106 | 74 | 69.8 | 32 | 30.2 |
| 1990 | 20 (1.7) | 94 (4.9) | 64 (4.1) | 2 (0.3) | 180 | 96 | 53.3 | 84 | 46.7 |
| 1991 | 3 (0.3) | 67 (2.9) | 18 (1.1) | 2 (0.3) | 90 | 40 | 44.4 | 50 | 55.6 |
| 1992 | 17 (1.4) | 151 (7.9) | 31 (2.0) | 1 (0.2) | 200 | 130 | 65.0 | 70 | 35.0 |
| 1993 | 34 (3.4) | 123 (6.5) | 34 (2.2) | 1 (0.2) | 192 | 131 | 68.2 | 61 | 31.8 |
| 1994 | 1 (0.1) | 10 (0.5) | 28 (1.8) | 5 (0.9) | 44 | 2 | 4.5 | 42 | 95.5 |
| 1995 | 0 (0.0) | 2 (0.1) | 3 (0.2) | 0 (0.0) | 5 | 0 | 0.0 | 5 | 100.0 |
| 1996 | 1 (0.1) | 33 (1.7) | 34 (2.2) | 0 (0.0) | 68 | 11 | 16.2 | 58 | 83.8 |
| 1997 | 2 (0.2) | 43 (2.3) | 27 (1.7) | 1 (0.2) | 73 | 30 | 41.1 | 43 | 58.9 |
| 1998 | 0 (0.0) | 3 (0.2) | 20 (1.3) | 3 (0.5) | 26 | 3 | 11.5 | 23 | 88.5 |
| 1999 | 1 (0.1) | 34 (1.8) | 6 (0.4) | 0 (0.0) | 41 | 3 | 7.3 | 38 | 92.7 |
| 2000 | 4 (0.4) | 68 (3.6) | 20 (1.3) | 0 (0.0) | 92 | 45 | 48.9 | 47 | 51.1 |
| 2001 | 24 (2.7) | 189 (9.9) | 84 (5.3) | 1 (0.2) | 298 | 168 | 56.4 | 130 | 43.6 |
| 2002 | 13 (1.4) | 227 (11.9) | 46 (2.9) | 13 (1.1) | 299 | 197 | 65.9 | 102 | 34.1 |
| Note: - indicates the river was not surveyed in that section during that year. |  |  |  |  |  |  |  |  |  |

## Genetic Sampling

No electrophoretic samples were collected from spring chinook recovered in the river or from the hatchery during spawning in 2002. We collected 192 DNA samples (opercle punches) from adult salmon ( 93 natural origin and 99 hatchery origin) and 207 samples from captive broodstock spawners. These samples were sent to the WDFW genetics lab in Olympia for analysis.

## Age Composition, Length Comparisons, and Fecundity

One objective of the monitoring program is to track the age composition of each year's returning adults. This allows us to annually compare ages of natural and hatchery-reared fish, and to examine long-term trends and variability in age structure. Overall, hatchery origin fish return at a younger age than natural origin fish (Figure 3). This difference is likely due to smolt size-atrelease (hatchery origin smolts are generally $25-30 \mathrm{~mm}$ greater in length than natural smolts).


Figure 3. Historical (1985-2001), and 2002 age composition for spring chinook in the Tucannon River.

Age at return for natural origin fish during 2002 was very similar to historical runs. Hatchery fish were composed of more Age 4 fish than historically observed. Age 5 hatchery-origin fish were absent from the 2002 run but are expected to increase substantially during the 2003 run due to the large run of Age 4 fish in 2002 and desirable ocean conditions.

Another comparison we conduct on returning adult natural and hatchery origin fish is the difference between mean post-eye to hypural-plate lengths. We reported in the past (Bumgarner et al. 1994) that hatchery fish were generally shorter than natural origin fish of the same age. For many of the early return years this appeared to be true (Figures 4, 5, 6, and 7). However, overall for all combined return years, there is no difference in mean length between natural and hatchery origin fish, even though they migrate as smolts at significantly different sizes (Bugert et al. 1990; Bugert et al. 1991).


Figure 4. Mean length and SD of Age 4 females.


Figure 6. Mean length and SD of Age 4 males.


Figure 5. Mean length and SD of Age 5 females.


Figure 7. Mean length and SD of Age 5 males.

Fecundities (number of eggs/female) of natural and hatchery origin fish from the Tucannon River program have been documented since 1990 (Table 8). Analysis of variance was performed to determine if there were significant differences in mean fecundities at the $95 \%$ confidence level. Natural origin females had significantly higher fecundities than hatchery origin fish for both Age $4(\mathrm{P}<0.001)$ and 5 -year-old fish $(\mathrm{P}<0.001)$.

Mean size of natural origin eggs in Age 4 spring chinook from the Tucannon River averaged $0.223 \mathrm{~g} / \mathrm{egg}$ and hatchery origin eggs averaged $0.239 \mathrm{~g} / \mathrm{egg}$. This difference was statistically significant at the $95 \%$ confidence level ( $\mathrm{P}<0.05$ ). This may help explain why Age 4 hatchery origin females are less fecund. However, mean egg size in Age 5 salmon was $0.269 \mathrm{~g} / \mathrm{egg}$ for natural origin and $0.270 \mathrm{~g} / \mathrm{egg}$ for hatchery origin females, but the difference was not significant ( $\mathrm{P}=0.95$ ).

Table 8. Average number of eggs/female (n, SD) by age group of Tucannon River natural and hatchery origin broodstock, 1990-2002.

| Year | Age 4 |  |  |  | Age 5 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Natural |  | Hatchery |  | Natural |  | Hatchery |  |
| 1990 | 3,691 | $(13,577.3)$ | 2,794 | $(18,708.0)$ | 4,383 | (8, 772.4) | No | Fish |
| 1991 | 2,803 | $(5,363.3)$ | 2,463 | ( 9, 600.8) | 4,252 | $(11,776.0)$ | 3,052 | $(1,000.0)$ |
| 1992 | 3,691 | $(16,588.3)$ | 3,126 | $(25,645.1)$ | 4,734 | (2, 992.8) | 3,456 | $(1,000.0)$ |
| 1993 | 3,180 | $(4,457.9)$ | 3,456 | $(5,615.4)$ | 4,470 | $(1,000.0)$ | 4,129 | $(1,000.0)$ |
| 1994 | 3,688 | $(13,733.9)$ | 3,280 | $(11,630.3)$ | 4,906 | (9, 902.0) | 3,352 | $(10,705.9)$ |
| 1995 | No | Fish | 3,584 | $(14,766.4)$ | 5,284 | $(6,136.1)$ | 3,889 | $(1,000.0)$ |
| 1996 | 3,509 | $(17,534.3)$ | 2,833 | $(18,502.3)$ | 3,617 | $(1,000.0)$ |  | Fish |
| 1997 | 3,487 | $(15,443.1)$ | 3,290 | $(24,923.3)$ | 4,326 | $(3,290.9)$ | No | Fish |
| 1998 | 4,204 | ( $1,000.0$ ) | 2,779 | $(7,375.4)$ | 4,017 | $(28,680.5)$ | 3,333 | $(6,585.2)$ |
| 1999 | No | Fish | 3,121 | $(34,445.4)$ | No | Fish | 3,850 | $(1,000.0)$ |
| 2000 | 4,144 | (2, 1,111.0) | 3,320 | $(34,545.4)$ | 3,618 | (1, 000.0) | 4,208 | $(1,000.0)$ |
| 2001 | 3,612 | $(27,508.4)$ | 3,225 | $(24,690.6)$ | No | Fish | 3,585 | (2, 842.5) |
| 2002 | 3,584 | $(14,740.7)$ | 3,368 | $(24,563.7)$ | 4,774 | $(7,429.1)$ |  | Fish |
| Mean |  | 3,596 |  | 3,185 |  | 4,376 |  | 474 |
| SD |  | 591.0 |  | 662.9 |  | 845.4 |  | 8.4 |

## Coded-Wire Tag Sampling

Broodstock collection, pre-spawn mortalities, and carcasses recovered from spawning ground surveys provide representatives of the annual run that can be sampled for CWT study groups (Table 9). Stray fish were predominately from the Umatilla River, Oregon and are discussed in more detail in a later section of this report. In 2002, based on the estimated escapement of fish to the river, we sampled approximately $31 \%$ of the run (Table 10).

Table 9. Coded-wire tag codes of hatchery salmon sampled at LFH and the Tucannon River, 2002.

| $\begin{array}{\|l\|l\|} \hline \text { CWT } \\ \text { Code } \end{array}$ | Broodstock Collected |  |  | Recovered in Tucannon River |  |  | Totals |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Died in Pond | $\begin{gathered} \text { Killed } \\ \text { Outright } \end{gathered}$ | Spawned | Dead in Trap | Pre-spawn <br> Mortality | Spawned |  |
| 63-12-11 | 2 |  | 47 |  | 1 | 104 | 154 |
| 63-02-75 |  |  | 2 |  |  |  | 2 |
| -Strays- |  |  |  |  |  |  |  |
| 10-54-12 |  |  |  |  |  | 1 | 1 |
| 07-61-38 |  |  |  |  |  | 1 | 1 |
| 07-60-51 |  |  |  |  |  | 1 | 1 |
| 07-60-49 |  |  |  |  |  | 1 | 1 |
| 07-60-41 |  |  |  |  |  | 2 |  |
| 07-60-40 |  |  | 1 |  |  | 1 | 2 |
| 07-60-39 |  |  | 1 |  |  |  | 1 |
| 05-42-08 |  |  |  |  |  | 1 | 1 |
| Lost tags |  |  |  |  |  | 2 | 2 |
| No tags |  | $10^{\text {a }}$ | 2 |  |  | 6 | 18 |
| Total | 2 | 10 | 53 | 0 | 1 | 120 | 186 |

Table 10. Spring chinook salmon (natural and hatchery) sampled from the Tucannon River, 2002.

|  | $\mathbf{2 0 0 2}$ |  |  |
| :--- | :---: | :---: | :---: |
|  | Natural | Hatchery | Total |
| Total escapement to river | 350 | 655 | 1,005 |
| Broodstock collected | 42 | 65 | 107 |
| Fish dead in adult trap | 0 | 0 | 0 |
| Total hatchery sample | 42 | 65 | 107 |
| Total fish left in river | 308 | 590 | 898 |
| In-river prespawn mortality | 0 | 1 | 1 |
| Spawned carcasses recovered | 81 | 120 | 201 |
| Total river sample | 81 | 121 | 202 |
| Carcasses sampled | 123 | 186 | 309 |

## Arrival and Spawn Timing Trends

Peak arrival and spawn timing have always been monitored to determine whether the hatchery program has caused a shift (Table 11). Peak arrival dates were based on greatest number of fish trapped on a single day. Peak spawn in the hatchery was determined by the day when the most females were spawned. Peak spawning in the river was determined by the highest weekly redd count.

Peak arrival during 2002 was at the average historical date for natural fish and earlier for hatchery fish as compared to previous years, but within the expected range compared to peak arrival before hatchery influence (1986-1988). Peak spawning date of hatchery fish in 2002 was also earlier than in previous years, although within the range found from previous years. The duration of active spawning in the Tucannon River was a week longer than previous years but that may be due in part to the larger run size.

Table 11. Peak dates of arrival of natural and hatchery salmon to the TFH adult trap and peak (date) and duration (number of days) for spawning in the hatchery and river, 1986-2002.

| Year | Peak Arrival at Trap |  | Spawning in Hatchery |  |  | Spawning in River |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Natural | Hatchery | Natural | Hatchery | Duration | Combined | Duration |
| 1986 | 5/27 | - | 9/17 | - | 31 | 9/16 | 36 |
| 1987 | 5/15 | - | 9/15 | - | 29 | 9/23 | 35 |
| 1988 | 5/24 | - | 9/07 | - | 22 | 9/17 | 35 |
| 1989 | 6/06 | 6/12 | 9/15 | 9/12 | 29 | 9/13 | 36 |
| 1990 | 5/22 | 5/23 | 9/04 | 9/11 | 36 | 9/12 | 42 |
| 1991 | 6/11 | 6/04 | 9/10 | 9/10 | 29 | 9/18 | 35 |
| 1992 | 5/18 | 5/21 | 9/15 | 9/08 | 28 | 9/09 | 44 |
| 1993 | 5/31 | 5/27 | 9/13 | 9/07 | 30 | 9/08 | 52 |
| 1994 | 5/25 | 5/27 | 9/13 | 9/13 | 22 | 9/15 | 29 |
| $1995{ }^{\text {a }}$ | - | 6/08 | 9/13 | 9/13 | 30 | 9/12 | 21 |
| 1996 | 6/06 | 6/20 | 9/17 | 9/10 | 21 | 9/18 | 35 |
| 1997 | 6/15 | 6/17 | 9/09 | 9/16 | 30 | 9/17 | 50 |
| 1998 | 6/03 | 6/16 | 9/08 | 9/16 | 36 | 9/17 | 16 |
| $1999{ }^{\text {a }}$ | - | 6/16 | 9/07 | 9/14 | 22 | 9/16 | 23 |
| 2000 | 6/06 | 5/22 | - | 9/05 | 22 | 9/13 | 30 |
| 2001 | 5/23 | 5/23 | 9/11 | 9/04 | 20 | 9/12 | 35 |
| Mean | 5/30 | 6/04 | 9/12 | 9/11 | 27 | 9/15 | 35 |
| 2002 | 5/29 | 5/29 | 9/10 | 9/03 | 22 | 9/11 | 42 |
| ${ }^{\text {a }}$ Too few natural salmon were trapped in 1995 and 1999 to determine peak arrival. |  |  |  |  |  |  |  |

## Total Run-Size

In general, redd counts have been directly related to total run-size entering the Tucannon River and passage of adult salmon at the TFH adult trap (Bugert et al. 1991). For 2002, we used sex ratios from collected broodstock and sex ratio observations on the spawning grounds to estimate the number of fish/redd. The run-size estimate for 2002 was calculated by adding the estimated number of fish upstream of the TFH adult trap, the estimated fish below the weir based on an estimated fish/redd ratio, the number of pre-spawn mortalities below the weir, and the number of broodstock collected (Table 12). Total run-size for 2002 was estimated at 1,005 fish ( 341 wild adults, 9 wild jacks and 644 hatchery-origin adults, 11 hatchery jacks). The total run for jacks and adults by origin has been estimated since 1985 (Appendix C).

Table 12. Estimated spring chinook salmon run to the Tucannon River, 1985-2002.

| Year ${ }^{\text {a }}$ | Total Redds | Fish/Redd Ratio ${ }^{\text {b }}$ | Spawning fish In the river | Broodstock Collected | Pre-spawning Mortalities ${ }^{\text {c }}$ | Total Run-Size | Percent <br> Natural |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 219 | 2.60 | 569 | 22 | 0 | 591 | 100 |
| 1986 | 200 | 2.60 | 520 | 116 | 0 | 636 | 100 |
| 1987 | 185 | 2.60 | 481 | 101 | 0 | 582 | 100 |
| 1988 | 117 | 2.60 | 304 | 125 | 0 | 429 | 96 |
| 1989 | 106 | 2.60 | 276 | 169 | 0 | 445 | 76 |
| 1990 | 180 | 3.39 | 611 | 135 | 8 | 754 | 66 |
| 1991 | 90 | 4.33 | 390 | 130 | 8 | 528 | 49 |
| 1992 | 200 | 2.82 | 564 | 97 | 92 | 753 | 56 |
| 1993 | 192 | 2.27 | 436 | 97 | 56 | 589 | 54 |
| 1994 | 44 | 1.59 | 70 | 70 | 0 | 140 | 70 |
| 1995 | 5 | 2.20 | 11 | 43 | 0 | 54 | 39 |
| 1996 | 68 | 2.00 | 136 | 80 | 16 | 232 | 63 |
| 1997 | 73 | 2.00 | 146 | 97 | 45 | 288 | 47 |
| 1998 | 26 | 1.94 | 51 | 89 | 4 | 144 | 59 |
| 1999 | 41 | 2.60 | 107 | 136 | 2 | 245 | 1 |
| 2000 | 92 | 2.60 | 239 | 81 | 19 | 339 | 24 |
| 2001 | 298 | 3.00 | 894 | 106 | 12 | 1,012 | 71 |
| 2002 | 299 | 3.00 | 897 | 107 | 1 | 1,005 | 35 |

${ }^{\text {a }}$ In 1994, 1995, 1998 and 1999, fish were not passed upstream, and in 1996 and 1997, high pre-spawning mortality occurred in fish passed above the trap, therefore; fish/redd ratio was based on the sex ratio of broodstock collected.
${ }^{\mathrm{b}}$ From 1985-1989 the TFH trap was temporary, thereby underestimating total fish passed upstream of the trap. The 1985-1989 fish/redd ratios were calculated from the 1990-1993 average, excluding 1991 because of a large jack run. ${ }^{\mathrm{c}}$ Effort in looking for pre-spawn mortalities has varied from year to year with more effort expended during years with poor conditions.

## Stray Salmon into the Tucannon River

Spring chinook from other river systems (strays) have periodically been recovered in the Tucannon River, though generally at a low proportion of the total run (Bumgarner et al. 2000). Through 1998 the incidence of stray spring chinook salmon was negligible (Table 13). However, in 1999, Umatilla River strays accounted for $8 \%$ of the total Tucannon River run, and that rate increased to $12 \%$ in 2000 . The increase in the number of strays, particularly from the Umatilla River, is a concern since it exceeds the allowable $5 \%$ stray rate of hatchery fish as deemed acceptable by NOAA Fisheries (formerly NMFS) and is contrary to WDFW management intent for the Tucannon River. Beginning with the 1997 brood year releases, the Oregon Department of Fish and Wildlife (ODFW) and Confederated Tribes of the Umatilla Indian Reservation (CTUIR) ceased marking a portion of Umatilla River origin spring chinook with an RV or LV fin clip (65-70\% of releases). Because of this action, fish that returned in 2002 were indistinguishable from wild origin spring chinook from the Tucannon River. For 2002, scale samples were collected from all wild fish collected for broodstock and passed upstream at the adult trap. Ten of the fish collected for broodstock (19\%) were determined to be of hatchery origin, based on scale pattern analysis, and were destroyed. Twelve of the unmarked fish sampled and passed upstream were found to be of hatchery origin (12\%) based on scale pattern analysis. Beginning with the 2000 BY, Umatilla River hatchery-origin spring chinook will be $100 \%$ marked. This will help ensure that Tucannon River spring chinook genetic integrity is maintained.

| Year | CWT <br> Code or Fin clip | Agency | Origin (stock) | Release Location / Release River | Number <br> Observed/ <br> Expanded ${ }^{\text {a }}$ | $\%$ of Tuc. Run |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | $\begin{aligned} & 074327 \\ & 074020 \\ & 232227 \\ & 232228 \end{aligned}$ | ODFW <br> ODFW <br> NMFS <br> NMFS | Carson (Wash.) <br> Rapid River <br> Mixed Col. <br> Mixed Col. | Meacham Cr. / Umatilla River | $2 / 5$ |  |
|  |  |  |  | Lookingglass Cr. / Grande Ronde | $1 / 2$ |  |
|  |  |  |  | Columbia River / McNary Dam | $2 / 5$ |  |
|  |  |  |  | Columbia River / McNary Dam | $1 / 2$ |  |
|  |  |  |  | Total Strays | 14 | 1.9 |
|  |  |  |  | Total Umatilla River | 5 | 0.7 |
| 1992 | $\begin{aligned} & 075107 \\ & 075111 \\ & 075063 \end{aligned}$ | ODFW <br> ODFW <br> ODFW | Lookingglass Cr . <br> Lookingglass Cr . <br> Lookingglass Cr . | Bonifer Pond / Columbia River | $2 / 6$ |  |
|  |  |  |  | Meacham Cr. / Umatilla River | $1 / 2$ |  |
|  |  |  |  | Meacham Cr. / Umatilla River | $1 / 2$ |  |
|  |  |  |  | Total Strays | 10 | 1.3 |
|  |  |  |  | Total Umatilla River | 4 | 0.5 |
| 1993 | 075110 | ODFW | Lookingglass Cr. | Meacham Cr. / Umatilla River | $1 / 2$ |  |
|  |  |  |  | Total Strays | 2 | 0.3 |
|  |  |  |  | Total Umatilla River | 2 | 0.3 |
| 1996 | $\begin{aligned} & 070251 \\ & \text { LV clip } \end{aligned}$ | ODFW <br> ODFW | Carson (Wash.) <br> Carson (Wash.) | Imeques AP / Umatilla River | $1 / 1$ |  |
|  |  |  |  | Imeques AP / Umatilla River | $1 / 2$ |  |
|  |  |  |  | Total Strays | 3 | 1.2 |
|  |  |  |  | Total Umatilla River | 3 | 1.2 |
| 1997 | $\begin{aligned} & 103042 \\ & 103518 \\ & \text { RV clip } \end{aligned}$ | IDFG <br> IDFG <br> ODFW | South Fork Salmon <br> Powell <br> Carson (Wash.) | Knox Bridge / South Fork Salmon | $1 / 2$ |  |
|  |  |  |  | Powell Rearing Ponds / Lochsa R. | $1 / 2$ |  |
|  |  |  |  | Imeques AP / Umatilla River | $3 / 5$ |  |
|  |  |  |  | Total Strays | 9 | 2.6 |
|  |  |  |  | Total Umatilla River | 5 | 1.4 |
| 1999 | 091751 | ODFW | Carson (Wash.) | Imeques AP / Umatilla River | $2 / 3$ |  |
|  | 092258 | ODFW | Carson (Wash.) | Imeques AP / Umatilla River | $1 / 1$ |  |
|  | 104626 | UI | Eagle Creek NFH | Eagle Creek NFH / Clackamas R. | $1 / 1$ |  |
|  | LV clip | ODFW | Carson (Wash.) | Imeques AP / Umatilla River | $2 / 2$ |  |
|  | RV clip | ODFW | Carson (Wash.) | Imeques AP / Umatilla River | $8 / 13$ |  |
|  |  |  |  | Total Strays | 20 | 8.2 |
|  |  |  |  | Total Umatilla River | 19 | 7.8 |

${ }^{\text {a }}$ All CWT codes recovered from groups that were $100 \%$ marked were given a $1: 1$ expansion rate. Groups that were not $100 \%$ marked were expanded based on the percentage of unmarked fish. The expansion is based on the percent of stray carcasses to Tucannon River origin carcasses and the estimated total run in the river.

| Table 13 (continued). Summary of identified stray hatchery origin spring chinook salmon that escaped into the Tucannon River (1990-2002). |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | CWT <br> Code or Fin clip | Agency | Origin (stock) | Release Location / Release River | Number <br> Observed/ <br> $\underset{\text { a }}{\text { Expanded }}$ | \% of Tuc. <br> Run |
| 2000 | $\begin{aligned} & 092259 \\ & 092260 \\ & 092262 \\ & 105137 \\ & 636330 \\ & 636321 \\ & \text { LV clip } \\ & \text { No Ad } \end{aligned}$ | ODFW <br> ODFW <br> ODFW <br> IDFG <br> WDFW <br> WDFW <br> ODFW <br> ODFW | Carson (Wash.) <br> Carson (Wash.) <br> Carson (Wash.) <br> Powell <br> Klickitat (Wash.) <br> Lyons Ferry (Wash.) <br> Carson (Wash.) <br> Carson (Wash.) | Imeques AP / Umatilla River Imeques AP / Umatilla River Imeques AP / Umatilla River Walton Creek/ Lochsa R. Klickitat Hatchery Lyons Ferry / Snake River Imeques AP / Umatilla River Imeques AP / Umatilla River Total Strays Total Umatilla River | $\begin{gathered} 4 / 4 \\ 1 / 1 \\ 1 / 3 \\ 1 / 3 \\ 1 / 1 \\ 1 / 1 \\ 18 / 31 \\ 2 / 2 \\ \mathbf{4 6} \\ \mathbf{4 1} \end{gathered}$ | $\begin{aligned} & 13.6 \\ & 12.1 \end{aligned}$ |
| 2001 |  | ODFW ODFW ODFW | Umatilla R. Imnaha R. \& Tribs. Imnaha R. \& Tribs. | Umatilla Hatch. /Umatilla River Lookinglass/Imnaha River Lookinglass/Imnaha River Total Strays Total Umatilla River | $\begin{gathered} 1 / 7 \\ 1 / 3 \\ 1 / 3 \\ \mathbf{1 3} \\ \mathbf{7} \end{gathered}$ | $\begin{aligned} & 1.3 \\ & 0.7 \end{aligned}$ |
| 2002 | $\begin{aligned} & \hline 054208 \\ & 076039 \\ & 076040 \\ & 076041 \\ & 076049 \\ & 076051 \\ & 076138 \\ & 105412 \end{aligned}$ | USFWS <br> ODFW <br> ODFW <br> ODFW <br> ODFW <br> ODFW <br> ODFW <br> IDFG | Dworshak Umatilla R. Umatilla R. Umatilla R. Umatilla R. Umatilla R. Umatilla R. Powell | Dworshak NFH/Clearwater River <br> Umatilla Hatch./Umatilla River Umatilla Hatch./Umatilla River Umatilla Hatch./Umatilla River Umatilla Hatch./Umatilla River Umatilla Hatch../ Umatilla River Umatilla Hatch./Umatilla River Clearwater Hatch./Powell Ponds Total Strays Total Umatilla River | $\begin{gathered} \hline 1 / 29 \\ 1 / 8 \\ 2 / 16 \\ 2 / 16 \\ 1 / 8 \\ 1 / 8 \\ 1 / 8 \\ 1 / 4 \\ \mathbf{9 7} \\ \mathbf{6 4} \end{gathered}$ | $\begin{aligned} & 9.6 \\ & 6.3 \end{aligned}$ |
| All CWT codes recovered from groups that were $100 \%$ marked were given a 1:1 expansion rate. Groups that were not $100 \%$ marked were expanded based on the percentage of unmarked fish. The expansion is based on the percent of stray carcasses to Tucannon River origin carcasses and the estimated total run in the river. |  |  |  |  |  |  |

## Juvenile Salmon Evaluation

## Hatchery Rearing, Marking, and Release

## Hatchery Rearing and Marking

Based on recommendations by Gallinat et al. (2001), the adipose clip was abandoned for Tucannon River spring chinook to prevent harvest of this listed population in sport fisheries. Supplementation juveniles (2001 BY) were marked with a red elastomer tag (VIE) behind the right eye and tagged with CWTs on September 17-30, 2002 (149,273 fish). Supplementation fish were transported to TFH on October 10. The 2001 BY captive brood juveniles (144,492 fish) were marked on September 9-13 with an agency-only wire tag in the snout and transported to TFH on October 11.
Length and weight samples were collected three times on the 2001 BY fish during the rearing cycle (Table 14). During February, fish were sampled for length, weight, hatchery mark quality, and PIT tagged for outmigration comparisons ( 1,010 supplementation fish and 1,007 captive brood progeny) before transfer to Curl Lake.

| Brood/ Date | Progeny Type | Sample Location | N | Mean Length | CV | K | FPP |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 |  |  |  |  |  |  |  |
| 5/03/02 | Supplementation | LFH | 200 | 69.0 | 6.6 | 1.11 | 123.4 |
| 2/14/03 | Supplementation | TFH | 250 | 125.5 | 15.5 | 1.19 | 18.2 |
| 4/08/03 | Supplementation | Curl Lake | 250 | 139.4 | 16.3 | 1.21 | 12.9 |
| 5/03/02 | Captive Brood | LFH | 200 | 62.5 | 9.7 | 1.09 | 165.0 |
| 2/18/03 | Captive Brood | TFH | 250 | 116.5 | 12.7 | 1.19 | 23.3 |
| 4/08/03 | Captive Brood | Curl Lake | 250 | 135.3 | 17.2 | 1.21 | 13.9 |

## 2001 Brood Release

A total of 21,043 marked (AD clip/CWT) excess supplementation parr (123 fish/lb) and 20,592 marked (AD clip/CWT) excess captive brood progeny ( 165 fish/lb) were released on May 6 at Bridge 11 (rkm 44) on the Tucannon River (Table 15). This parr release allowed WDFW to stay within the maximum allowed number of smolts released under Section 10 Permit \#1129 ( 150,000 supplementation, 150,000 captive brood progeny). The 2001 BY pre-smolts were transported to Curl Lake in February 2003 for acclimation and volitional release.

Volitional release began April 1 and continued until April 21 when remaining fish were forced out. Mortalities were low in Curl Lake and WDFW released an estimated 146,922
supplementation fish ( 12.9 fish/lb) and 140,396 captive broodstock progeny ( 13.9 fish/lb).

| Release <br> Year | (BY) | Release |  | CWT <br> Code | Number CWT | AD-onlymarked | Additional mark/cross | Ibs | $\begin{gathered} \text { Fish/ } \\ \text { lb } \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Location | Date |  |  |  |  |  |  |
| 2002 | (01) | Bridge 11 | 5/06 | 63/14/29 | 19,948 | 1,095 | No VI, Mixed | 170.5 | 123.4 |
| 2002 | (01CB | Bridge 11 | 5/06 | 63/14/30 | 20,435 | 157 | No VI, Mixed | 124.8 | 165.0 |
|  | ) |  |  |  |  |  |  |  |  |
| 2003 |  | Curl Lake | 4/01-4/21 | 63/06/81 | 146,922 | N/A | Rt. Red VI, Mixed | 11,389 | 12.9 |
| 2003 | (01) | Curl Lake | 4/01-4/21 | 63 | 140,396 | N/A | No VI, Mixed | 10,100 | 13.9 |
|  | $(01 \mathrm{CB}$ |  |  |  |  |  |  |  |  |

N/A = Not applicable.

## Natural Parr Production

Program staff surveyed the Tucannon River at index sites in 2002 to estimate the density and population of subyearling (Table 16, Appendix D) and yearling spring chinook salmon. Snorkel surveys were conducted using a total count method (Griffith 1981, Schill and Griffith 1984). Population size was determined by multiplying the mean fish density (fish $/ 100 \mathrm{~m}^{2}$ ) by the estimated total area within each stratum. Fifty sites were snorkeled in 2002 (August 12-13). Total area snorkeled was approximately $5.0 \%$ of the suitable rearing habitat in the Tucannon River. A total of 3,017 subyearling and 35 yearling spring chinook were counted during the surveys. We estimated that $63,412( \pm 11,733)$ subyearling and $703( \pm 374)$ yearling chinook were present in the river (Table 16).

Table 16. Number of sites, area snorkeled, mean density (fish $/ 100 \mathrm{~m}^{2}$ ), population estimates, and $95 \%$ confidence intervals for subyearling and yearling spring chinook within the Tucannon River, 2002.

| Stratum | Number of sites | Area ( $\mathrm{m}^{2}$ ) snorkeled | Subyearling |  |  | Yearling |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Mean Density | Pop. <br> Estimate | C.I. | Mean Density | Pop. <br> Estimate | C.I. |
| Marengo | 6 | 2,973 | 9.79 | 5,625 | 2,823 | 0.04 | 23 | 44 |
| Hartsock | 14 | 7,921 | 14.01 | 24,735 | 6,864 | 0.04 | 68 | 76 |
| HMA | 20 | 11,771 | 12.60 | 28,175 | 7,926 | 0.16 | 350 | 283 |
| Wilderness | 10 | 4,155 | 6.05 | 4,877 | 4,503 | 0.32 | 262 | 198 |
| Total | 50 | 26,820 | 11.35 | 63,412 | 11,733 | 0.14 | 703 | 374 |

## Natural Smolt Production

Program staff operated a 1.5 m rotary screw trap at rkm 3 on the Tucannon River from October 14,2001 to July 15,2002 to estimate numbers of migrating natural and hatchery spring chinook. The smolt trap was pulled from October 23-January 22 to convert the smolt trap drum cleaning system to a paddle wheel drive design. The trap was pulled again from April 6-10 to fix a hole in the cone. Other data on natural and hatchery spring chinook smolts such as peak outmigration, length of smolts, etc., have not been reported here for simplicity. Those data are available upon request.

We examined the influence of specific abiotic variables on spring chinook emigration during the last five trapping seasons (1997/1998 to 2001/2002) using correlation analysis. Significant relationships were found between the total number of wild spring chinook smolts captured ( $\log _{10}$ transformed for normality) emigrating from the Tucannon River and flow $\left(\mathrm{ft}^{3} / \mathrm{sec}\right)\left(\mathrm{r}^{2}=0.26, \mathrm{P}<\right.$ 0.01 ), staff gauge level ( $\mathrm{r}^{2}=0.21, \mathrm{P}<0.01$ ), time of year ( $\mathrm{r}^{2}=0.27, \mathrm{P}<0.01$ ), water temperature $\left(\mathrm{r}^{2}=0.24, \mathrm{P}<0.01\right)$, and secchi reading ( $\mathrm{r}^{2}=0.28, \mathrm{P}<0.01$ ). Although these variables are statistically significant, they account for only a small amount of the variability in the number of emigrating fish. This is understandable as smoltification is a physiological process and the resulting outmigration may only be slightly influenced by abiotic factors. No significant relationships were found between the numbers of hatchery spring chinook smolts captured ( $\log _{10}$ transformed) and flow, staff gauge level, time of year, water temperature, or secchi disk reading.

Each week we attempted to determine trap efficiency by clipping a portion of the caudal fin on a few representative captured migrants and releasing them one kilometer upstream. The percent of marked fish recaptured was used as an estimate of weekly trapping efficiency. To calculate trapping efficiency during weeks when low numbers of fish were caught we examined the relationship between trap efficiency and the variables flow, staff gauge, number of fish captured, water temperature, and time of year (week). There were no statistically significant relationships between trap efficiency for wild and hatchery spring chinook and any of the variables examined. Despite the low statistical power, we believe that trap efficiency decreases as flow increases.

Flow is the dominant factor affecting downstream migrant trapping operations in any system according to Seiler et al. (1999). Groot and Margolis (1991) state that the rate of downstream migration of chinook fingerlings appears to be both time and size dependent and may also be related to river discharge and the location of fish in the river. They state that during years of low and stable river flow; the rate of downstream migration was negatively correlated with discharge, whereas, when flows were higher and more variable, the rate of migration was positively correlated with discharge.

Mean daily flow data was provided by the U.S. Geological Survey gauge station at Starbuck, WA (rkm 12.7). Correlation analysis indicated a statistically significant relationship between
flow and the staff gauge level at the smolt trap at the $99 \%$ confidence level $\left(\mathrm{r}^{2}=0.85\right)$. As the U.S.G.S. flow data is computer monitored on a continuous basis and is in relatively close proximity to the smolt trap, we estimated trap efficiencies based on the U.S.G.S. flow data with the following equations:

## Wild Spring Chinook

Trap Efficiency $=22.48-0.01$ (Flow)

## Hatchery Spring Chinook <br> Trap Efficiency $=22.09-0.04$ (Flow)

To estimate potential juvenile migrants passing when the trap was not operated for short intervals, such as periods when freshets washed out large amounts of debris from the river, we calculated the average number of fish trapped for three days before and three days after nontrapping periods. The mean number of fish trapped daily was then divided by the estimated trap efficiency to calculate fish passage. The estimated number of fish passing each day was then applied to each day the trap was not operated.

During the October 23 - January 22 time period when the smolt trap was being modified the number of wild spring chinook that would have been captured was predicted by an exponential model based on the 2000 and 2001 smolt trapping data using the following equation:

Number of wild spring chinook $=\exp [0.114+0.014$ (Flow) $] \quad\left(\mathrm{r}^{2}=0.47 ; \mathrm{P}<0.01\right)$
We estimated that 20,049 , or $45 \%$ of the 2000 BY parr estimates, passed the smolt trap during 2001-2002 (Table 17). We also estimated that $90 \%$ of the hatchery supplementation fish and $32 \%$ of the captive brood progeny released from Curl Lake Acclimation Pond (2000 BY) passed the smolt trap.

Table 17. Monthly and total population estimates, with $95 \%$ confidence intervals, for natural and hatchery origin (supplementation and captive brood) emigrants from the Tucannon River, 2002.

| Month | Natural | $+/-\mathbf{9 5 \%}$ C.I. | Hatchery | $+/ \mathbf{9 5 \%}$ C.I. | C.B. ${ }^{\text {b }}$ | $+/ \mathbf{9 5 \%}$ <br> C.I. |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Sept.-Feb. | 496 | 26 | 0 | -- | 0 | - |
| March | 367 | 27 | 0 | - | 0 | -- |
| April | 11,375 | 884 | 13,957 | 2,819 | 266 | 44 |
| May | 7,753 | 533 | 77,949 | 8,345 | 697 | 104 |
| June | 58 | 10 | 301 | 32 | 0 | 0 |
| Total | 20,049 | 1,480 | 92,207 | 11,196 | 963 | 148 |
| \% Survival $^{\text {a }}$ | 44.9 |  | 90.3 |  | 31.5 |  |

${ }^{\text {a }}$ Percent survival to smolt based on estimated number of parr from summer snorkel surveys (natural origin) or from TFH release numbers (hatchery origin).
${ }^{\mathrm{b}}$ C.B. $=$ captive brood progeny.

## Juvenile Migration Studies

In 2002, WDFW used Passive Integrated Transponder (PIT) tags to study the emigration timing and success of wild and hatchery origin spring chinook. The tags allowed us to identify the characteristics of successful smolts. We tagged 321 wild and 318 -hatchery origin spring chinook over a two-week period (Table 18). No fish were killed during PIT tagging, though it is likely that some delayed mortality occurred after release. Detection rates were generally higher for wild chinook and mean travel days were generally higher for hatchery spring chinook. Detection rates may be higher for wild chinook because they are smaller (25-48 mm less) and more likely to be captured at collection facilities, or survival rates were actually slightly higher.

Table 18. Cumulative detection (one unique detection per tag code) and travel time (TD) summaries of PIT tagged spring chinook salmon released from the Tucannon River smolt trap (rkm 3) at downstream Snake and Columbia River dams in 2002.

| Release <br> Date | Release Data |  | Mean Length | Mean <br> Length | Recapture Data |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | LMJ |  | MCJ |  | JDJ |  | BONN |  | Total |
|  | Origin | N |  |  | N | TD | N | TD | N | TD | N | TD | N (\%) |
| 4/29-02 | W | 220 |  | 105.5 | 105.6 | 93 | 3.8 | 41 | 10.1 | 18 | 14.1 | 7 | 18.9 | 159 (72.3) |
|  | H | 219 | 138.2 | 139.8 | 77 | 5.4 | 28 | 11.6 | 12 | 16.4 | 6 | 22.1 | 123 (56.2) |
| 5/06-08 | W | 101 | 105.3 | 105.0 | 31 | 3.3 | 21 | 8.9 | 6 | 14.1 | 4 | 15.4 | 62 (61.4) |
|  | H | 99 | 138.0 | 136.5 | 29 | 5.2 | 24 | 10.3 | 5 | 14.7 | 4 | 15.5 | 62 (62.6) |

Note: Mean travel times listed are from the total number of fish detected at each dam, not just unique recoveries for a tag code. Abbreviations are as follows: LMJ-Lower Monumental Dam, MCJ- McNary Dam, JDJ-John Day Dam, BONN-Bonneville Dam, TD- Mean Travel Days.

Survival probabilities were estimated by the Cormack Jolly-Seber methodology using the Survival Under Proportional Hazards (SURPH2) computer model. The data files were created using the CAPTHIST program. Data for input into CAPHIST was obtained directly from PTAGIS. Survival estimates to Lower Monumental Dam were $0.94( \pm 0.06)$ and $0.83( \pm 0.07)$ for wild and hatchery-origin fish, respectively. While survival estimates were slightly lower for hatchery-origin fish the differences were not significant. Survival estimates to John Day Dam were nearly identical at $0.67( \pm 0.15)$ for wild and $0.66( \pm 0.25)$ for hatchery-origin fish.

## Survival Rates

Point estimates of population sizes have been calculated for various life stages (Tables 19 and 20) of natural origin fish from spawning ground and juvenile mid-summer population surveys, smolt trapping, and fecundity estimates. From these two tables, survivals between life stages have been calculated for both natural and hatchery salmon to assist in the evaluation of the hatchery program. These survival estimates provide insight as to where efforts should be directed to improve not only the survival of fish produced within the hatchery, but fish in the river as well.

As expected, juvenile (egg-fry-smolt) survival rates for hatchery fish are considerably higher than for naturally reared salmon (Table 21) because they have been protected in the hatchery. However, smolt-to-adult return rates (SAR) of natural salmon were about three times higher than for hatchery-reared salmon (Tables 22 and 23). The mean hatchery SAR's ( $0.15 \%$ ) documented from the 1985-1997 broods were below the goal SAR of $0.87 \%$ established under the LSRCP. Hatchery SAR's for Tucannon River salmon need to substantially improve to meet the mitigation goal of 1,152 salmon.

Table 19. Estimates of natural Tucannon spring chinook salmon abundance by life stage for 1985-2002 broods.

| Brood Year | Females in river |  | Mean ${ }^{\text {a fecundity }}$ |  | $\begin{gathered} \text { Number of } \\ \text { eggs } \end{gathered}$ | $\begin{gathered} \text { Number }{ }^{\text {b }} \text { of } \\ \text { fry } \\ \hline \end{gathered}$ | $\begin{gathered} \text { Number of } \\ \text { smolts } \end{gathered}$ | Progeny ${ }^{\text {c }}$ (returning adults) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | natural | hatchery | natural | hatchery |  |  |  |  |
| 1985 | 219 | - | 3,883 | - | 850,377 | 90,200 | 42,000 | 392 |
| 1986 | 200 | - | 3,916 | - | 783,200 | 102,600 | 58,200 | 468 |
| 1987 | 185 | - | 4,096 | - | 757,760 | 79,100 | 44,000 | 238 |
| 1988 | 117 | - | 3,882 | - | 454,194 | 69,100 | 37,500 | 527 |
| 1989 | 103 | 3 | 3,883 | 2,606 | 407,767 | 58,600 | 30,000 | 158 |
| 1990 | 128 | 52 | 3,993 | 2,697 | 651,348 | 86,259 | 49,500 | 94 |
| 1991 | 51 | 39 | 3,741 | 2,517 | 288,954 | 54,800 | 30,000 | 7 |
| 1992 | 119 | 81 | 3,854 | 3,295 | 725,521 | 103,292 | 50,800 | 194 |
| 1993 | 112 | 80 | 3,701 | 3,237 | 673,472 | 86,755 | 49,560 | 204 |
| 1994 | 39 | 5 | 4,187 | 3,314 | 179,863 | 12,720 | 7,000 | 12 |
| 1995 | 5 | 0 | 5,224 | 0 | 26,120 | 0 | 75 | 6 |
| 1996 | 53 | 16 | 3,516 | 2,843 | 231,836 | 2,845 | 1,612 | 69 |
| 1997 | 39 | 33 | 3,609 | 3,315 | 250,146 | 32,913 | 21,057 | 803 |
| 1998 | 19 | 7 | 4,023 | 3,035 | 97,682 | 8,453 | 5,508 | 266 |
| 1999 | 1 | 40 | 3,965 | 3,142 | 129,645 | 15,944 | 8,157 | 9 |
| 2000 | 26 | 66 | 3,969 | 3,345 | 323,964 | 44,618 | 20,045 |  |
| 2001 | 219 | 79 | 3,612 | 3,252 | 1,047,936 | 63,412 |  |  |
| 2002 | 104 | 195 | 3,981 | 3,368 | 1,070,784 |  |  |  |

a 1985 and 1989 mean fecundity of natural females is average of 1986-88 and 1990-93.
b Number of fry estimated from electrofishing (1985-1989), Line transect snorkel surveys (1990-1992), and Total Count snorkel surveys (1993-1999).
c Numbers do not include down river harvest estimates or out-of-basin recoveries.

Table 20. Estimates of Tucannon spring chinook salmon abundance (spawned and reared in the hatchery) by life stage for 1985-2002 broods.

| Brood Year | Females spawned |  | Mean ${ }^{\text {a }}$ fecundity |  | Number of eggs | Number of fry | Number of smolts | $\begin{gathered} \text { Progeny }{ }^{\text {b }} \\ \text { (returning } \\ \text { adults) } \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | natural | hatchery | natural | hatchery |  |  |  |  |
| 1985 | 4 | - | 3,883 | - | 14,843 | 13,401 | 12,922 | 45 |
| 1986 | 57 | - | 3,916 | - | 187,958 | 177,277 | 153,725 | 339 |
| 1987 | 48 | - | 4,096 | - | 196,573 | 164,630 | 152,165 | 190 |
| 1988 | 49 | - | 3,882 | - | 182,438 | 150,677 | 146,200 | 447 |
| 1989 | 28 | 9 | 3,883 | 2,606 | 133,521 | 103,420 | 99,060 | 243 |
| 1990 | 21 | 23 | 3,993 | 2,697 | 126,334 | 89,519 | 85,800 | 28 |
| 1991 | 17 | 11 | 3,741 | 2,517 | 91,275 | 77,232 | 74,060 | 25 |
| 1992 | 28 | 18 | 3,854 | 3,295 | 156,359 | 151,727 | $87,752^{\text {c }}$ | 81 |
| 1993 | 21 | 28 | 3,701 | 3,237 | 168,366 | 145,303 | 138,848 | 207 |
| 1994 | 22 | 21 | 4,187 | 3,314 | 161,707 | 132,870 | 130,069 | 34 |
| 1995 | 6 | 15 | 5,224 | 0 | 85,772 | 63,935 | 62,272 | 180 |
| 1996 | 18 | 19 | 3,516 | 2,843 | 117,287 | 80,325 | 76,219 | 260 |
| 1997 | 17 | 25 | 3,609 | 3,315 | 144,237 | 29,650 | 24,184 | 181 |
| 1998 | 30 | 14 | 4,023 | 3,035 | 161,019 | 136,027 | 127,939 | 666 |
| 1999 | 1 | 36 | 3,965 | 3,142 | 113,544 | 106,880 | 97,600 | 7 |
| 2000 | 3 | 35 | 3,969 | 3,345 | 128,980 | 123,313 | 102,099 |  |
| 2001 | 29 | 27 | 3,612 | 3,252 | 184,127 | 174,934 | 146,922 |  |
| 2002 | 22 | 25 | 3,981 | 3,368 | 169,364 | 151,531 |  |  |
| ${ }^{\text {a }} 1985$ and 1989 mean fecundity of natural females is average of 1986-88 and 1990-93, 1999 mean fecundity natural fish is the based on the mean of 1986-1998. <br> ${ }^{b}$ Numbers do not include down river harvest estimates or out of basin recoveries. <br> ${ }^{\text {c }}$ Number of smolts is less than actual release number. 57,316 parr were released in October 1993, with an estimated $7 \%$ survival. Total number of hatchery fish released from the 1992 brood year was 140,725 . We therefore use the listed number of 87,752 as the number of smolts released. |  |  |  |  |  |  |  |  |


| Brood Year | Natural |  |  | Hatchery |  |  | Hatchery Advantage |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Egg to fry | Fry to smolt | Egg to smolt | $\begin{gathered} \text { Egg to } \\ \text { fry } \\ \hline \end{gathered}$ | Fry to smolt | Egg to smolt | Egg to fry | Fry to smolt | Egg to smolt |
| 1985 | 10.6 | 46.6 | 4.9 | 90.3 | 96.4 | 87.1 | 8.5 | 2.1 | 17.6 |
| 1986 | 13.1 | 56.7 | 7.4 | 94.3 | 86.7 | 81.8 | 7.2 | 1.5 | 11.0 |
| 1987 | 10.4 | 55.6 | 5.8 | 83.8 | 92.4 | 77.4 | 8.0 | 1.7 | 13.3 |
| 1988 | 15.2 | 54.3 | 8.3 | 82.6 | 97.0 | 80.1 | 5.4 | 1.8 | 9.7 |
| 1989 | 14.4 | 51.2 | 7.4 | 77.5 | 95.8 | 74.2 | 5.4 | 1.9 | 10.1 |
| 1990 | 13.2 | 57.4 | 7.6 | 70.9 | 95.8 | 67.9 | 5.4 | 1.7 | 8.9 |
| 1991 | 19.0 | 54.7 | 10.4 | 84.6 | 95.9 | 81.1 | 4.5 | 1.8 | 7.8 |
| 1992 | 14.2 | 49.2 | 7.0 | 97.0 | 57.8 | 56.1 | 6.8 | 1.2 | 8.0 |
| 1993 | 12.9 | 57.1 | 7.4 | 86.3 | 95.6 | 82.5 | 6.7 | 1.7 | 11.2 |
| 1994 | 7.1 | 55.0 | 3.9 | 82.2 | 97.9 | 80.4 | 11.6 | 1.8 | 20.7 |
| 1995 | 0.0 | 0.0 | 0.3 | 74.5 | 97.4 | 72.6 | -- | -- | -- |
| 1996 | 1.2 | 56.7 | 0.7 | 68.5 | 94.9 | 65.0 | 55.8 | 1.7 | -- |
| 1997 | 13.2 | 64.0 | 8.4 | 20.6 | 81.6 | 16.8 | 1.6 | 1.3 | 2.0 |
| 1998 | 8.7 | 65.2 | 5.6 | 84.5 | 94.1 | 79.5 | 9.8 | 1.4 | 14.1 |
| 1999 | 12.3 | 51.2 | 6.3 | 94.1 | 91.3 | 86.0 | 7.7 | 1.8 | 13.7 |
| 2000 | 13.8 | 44.9 | 6.2 | 95.6 | 82.8 | 79.2 | 6.9 | 1.8 | 12.8 |
| 2001 | 6.1 |  |  | 95.0 | 84.0 | 79.8 | 15.7 |  |  |
| 2002 |  |  |  | 89.5 |  |  |  |  |  |
| Mean | 10.9 | 51.2 | 6.1 | 81.8 | 90.4 | 73.4 | 10.4 | 1.7 | 11.5 |
| SD | 5.0 | 14.7 | 2.7 | 17.5 | 10.0 | 16.6 | 12.5 | 0.2 | 4.5 |


|  |  | Num | of | Retu | erv |  | exp) ${ }^{\text {a }}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  | SAR | (\%) |
| Year | smolts | obs | $\exp$ | obs | exp | obs | exp | w/ jacks | no jacks |
| 1985 | 42,000 | 8 | 19 | 110 | 255 | 36 | 118 | 0.93 | 0.89 |
| $1986{ }^{\text {b }}$ | 58,200 | 1 | 2 | 115 | 376 | 28 | 90 | 0.80 | 0.80 |
| 1987 | 44,000 | 0 | 0 | 52 | 167 | 29 | 71 | 0.54 | 0.54 |
| 1988 | 37,500 | 1 | 3 | 136 | 335 | 74 | 189 | 1.41 | 1.40 |
| 1989 | 30,000 | 5 | 12 | 47 | 120 | 23 | 26 | 0.53 | 0.49 |
| 1990 | 49,500 | 3 | 8 | 63 | 72 | 12 | 14 | 0.19 | 0.17 |
| 1991 | 30,000 | 0 | 0 | 4 | 5 | 1 | 2 | 0.02 | 0.02 |
| 1992 | 50,800 | 2 | 2 | 84 | 159 | 16 | 33 | 0.38 | 0.38 |
| 1993 | 49,560 | 1 | 2 | 62 | 127 | 58 | 75 | 0.41 | 0.41 |
| 1994 | 6,000 | 0 | 0 | 8 | 10 | 1 | 2 | 0.20 | 0.20 |
| 1995 | 75 | 0 | 0 | 1 | 1 | 2 | 5 | $8.00{ }^{\text {c }}$ | $8.00^{\text {c }}$ |
| 1996 | 1,612 | 0 | 0 | 27 | 63 | 2 | 6 | 4.28 | 4.28 |
| 1997 | 21,057 | 6 | 14 | 234 | 703 | 29 | 86 | 3.81 | 3.75 |
| Geometric Mean of 1985-1997 broods |  |  |  |  |  |  |  | 0.54 | 0.53 |

${ }^{a}$ Expanded numbers are calculated from the proportion of each known age salmon recovered in the river and from broodstock collections in relation to the total estimated return to the Tucannon River. Expansions do not include down river harvest or Tucannon River fish straying to other systems.
${ }^{\mathrm{b}}$ One known (expanded to two) age 6 salmon was recovered.
${ }^{c} 1995$ SAR not included in mean.

Table 23. Adult returns and SAR's of hatchery salmon to the Tucannon River for brood years 1985-1997.

|  |  | Num | of A | Return | nown | expan | (exp.) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Ag |  |  |  |  |  | SAR | (\%) |
| Year | smolts | known | exp. | known | exp. | known | exp. | w/ jacks | no jacks |
| 1985 | 12,922 | 9 | 19 | 25 | 26 | 0 | 0 | 0.35 | 0.20 |
| 1986 | 153,725 | 79 | 83 | 99 | 238 | 8 | 18 | 0.22 | 0.17 |
| 1987 | 152,165 | 9 | 22 | 70 | 151 | 8 | 17 | 0.12 | 0.11 |
| 1988 | 146,200 | 46 | 99 | 140 | 295 | 26 | 53 | 0.31 | 0.24 |
| 1989 | 99,057 | 7 | 15 | 100 | 211 | 14 | 17 | 0.25 | 0.23 |
| 1990 | 85,500 | 3 | 6 | 16 | 20 | 2 | 2 | 0.03 | 0.03 |
| 1991 | 74,058 | 4 | 5 | 20 | 20 | 0 | 0 | 0.03 | 0.03 |
| 1992 | 87,752 | 11 | 11 | 50 | 66 | 2 | 4 | 0.09 | 0.08 |
| 1993 | 138,848 | 11 | 15 | 93 | 174 | 15 | 18 | 0.15 | 0.14 |
| 1994 | 130,069 | 2 | 4 | 21 | 25 | 4 | 5 | 0.03 | 0.02 |
| 1995 | 62,272 | 13 | 16 | 117 | 160 | 2 | 4 | 0.29 | 0.26 |
| 1996 | 76,219 | 44 | 60 | 100 | 186 | 5 | 14 | 0.34 | 0.26 |
| 1997 | 24,186 | 7 | 13 | 59 | 168 | 0 | 0 | 0.75 | 0.69 |
| Geometric Mean of 1985-1997 broods |  |  |  |  |  |  |  | 0.15 | 0.13 |

We found a significant relationship between survival calculated from CWT returns through the Regional Mark Information System (RMIS) database and size of smolts at release, with larger fish (6-10 fish/lb) having higher survival ( $\mathrm{r}^{2}=31.3, \mathrm{P}<0.05$ ) (Appendix E). However, years in which smaller fish (14-19 fish/lb) were released also coincided with poor ocean conditions, drought years, and flood events within the Tucannon River watershed. Decreasing the release size of smolts has allowed hatchery fish to more closely resemble wild fish and decrease the incidence of precocious fish and returning jacks, but overall survival appears to have decreased. An experimental release of fish at $15 / \mathrm{lb}$ and $10 / \mathrm{lb}$ during the same year would provide a direct comparison of differences in survival, age structure, length, and fecundity of adult returns.

While SAR's were lower for hatchery salmon, overall survival of hatchery salmon to return as adults was higher than naturally reared fish because of the early-life survival advantage provided
by the hatchery (Table 21). With the exception of the 1988, 1997 and 1998 brood years, naturally produced fish have been below the replacement level (Figure 8; Table 24). Based on adult returns from the 1985-1997 broods, naturally reared salmon produced 0.5 adults for every spawner, while hatchery reared fish produced 2.0 adults.


Figure 8. Return per spawner ratio (with replacement line) for the 1985-1998 brood years.

Table 24. Parent-to-progeny survival estimates of Tucannon River spring chinook salmon from 1985 through 1998 brood years (1998 incomplete).

|  | Natural Salmon |  |  |  |  | Hatcherv Salmon |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Brood <br> Year | Number of <br> spawners | Number of <br> returns | Return/ <br> spawner | Number of <br> spawners | Number of <br> returns | Return/ <br> Rpawner | Natural <br> advantage |  |  |  |
| 1985 | 569 | 392 | 0.69 | 9 | 45 | 5.00 | 7.2 |  |  |  |
| 1986 | 520 | 468 | 0.90 | 91 | 339 | 3.73 | 4.1 |  |  |  |
| 1987 | 481 | 238 | 0.49 | 83 | 190 | 2.29 | 4.7 |  |  |  |
| 1988 | 304 | 527 | 1.73 | 87 | 447 | 5.14 | 3.0 |  |  |  |
| 1989 | 276 | 158 | 0.57 | 122 | 243 | 1.99 | 3.5 |  |  |  |
| 1990 | 611 | 94 | 0.15 | 78 | 28 | 0.36 | 2.4 |  |  |  |
| 1991 | 390 | 7 | 0.02 | 72 | 25 | 0.35 | 17.5 |  |  |  |
| 1992 | 564 | 194 | 0.34 | 83 | 81 | 0.98 | 2.9 |  |  |  |
| 1993 | 436 | 204 | 0.47 | 91 | 207 | 2.27 | 4.8 |  |  |  |
| 1994 | 70 | 12 | 0.17 | 69 | 34 | 0.49 | 2.9 |  |  |  |
| 1995 | 11 | 6 | 0.55 | 39 | 180 | 4.62 | 8.4 |  |  |  |
| 1996 | 138 | 69 | 0.51 | 74 | 260 | 3.51 | 6.9 |  |  |  |
| 1997 | 146 | 803 | 5.50 | 89 | 181 | 2.03 | 0.4 |  |  |  |
| 1998 | 51 | 266 | 5.22 | 85 | 666 | 7.84 | 1.5 |  |  |  |
| Geometric |  |  |  |  |  |  |  |  |  |  |
| Mean |  |  | $\mathbf{0 . 5 4}$ |  |  | $\mathbf{1 . 9 9}$ | $\mathbf{3 . 7}$ |  |  |  |

## Fishery Contribution

An original goal of the LSRCP supplementation program was to enhance wild (natural) returns of salmon to the Tucannon River by providing 1,152 hatchery-reared fish (the number estimated to have been lost due to the construction of the Lower Snake River hydropower system) to the river. Such an increase would allow for limited harvest of the stock and increased spawning. However, hatchery adult returns have always been below the program goal. Moreover, natural escapement, with the exception of the 2001 and 2002 runs, has been low (Figure 9). Based on 1985-1997 brood year CWT recoveries from the RMIS database (Appendix E), sport and commercial harvest combined has only accounted for approximately $5.9 \%$ of the hatchery adult fish recovered annually. While exploitation has been relatively low, fishing mortality is one form of mortality fisheries managers can control. Adipose clipped hatchery fish have traditionally been targeted in the sport fishery. This hatchery fin clip was abandoned for Tucannon River spring chinook smolts starting with the 2000 BY to mitigate fishing mortality on this ESA listed population. Supplementation fish are now marked with a CWT and a red VIE tag behind the right eye. Captive brood progeny are marked only with agency-only wire tags to distinguish them from supplementation origin fish. Out-of-basin stray rates of Tucannon River spring chinook have been low (Appendix E), with an average of $3.7 \%$ of the adult hatchery fish straying to other river systems/hatcheries for brood years 1985-1997 (range 0-20\%).


Figure 9. Total escapement for Tucannon River spring chinook salmon for the 1985-2002 run years.

## Conclusions and Recommendations

Washington's LSRCP hatchery spring chinook salmon program has failed to return adequate numbers of adults to meet the mitigation goal of the program. This occurred because SARs of hatchery origin fish have consistently been below the assumed SAR of hatchery smolts as described under the LSRCP, even though hatchery returns have generally been at 2-3 times the replacement level. Further, the natural population of spring chinook salmon in the river has declined and remained below the replacement level for most years, with the majority ( $95 \%$ ) of the mortality occurring between the green egg and smolt stages. Ocean conditions and mortality within the mainstem migration corridor have also contributed to the decline. The end result has been a slow but steady replacement of the natural population with the hatchery population. While this neither was, nor is the desired result of the program, in many ways the hatchery program has helped conserve the natural population within the river by returning enough adults to allow some spawning in the river. System survivals (in-river, ocean) must increase in the future for the hatchery program and the natural run to reach its full potential, and the spring chinook run returned to its historic levels.

Until that time, the evaluation program will continue to document and study life history survivals, genotypic and phenotypic traits, and examine procedures within the hatchery that can be improved to benefit the program and the natural population. Based on our previous studies and current data involving survival and physical characteristics we recommend the following:

1. We continue to see annual differences in phenotypic characteristics of returning salmon (i.e.,
hatchery fish are generally younger in age and less fecund than natural origin fish), yet other traits such as run and spawn time have changed little over the program's history. Further, genetic analysis to date indicates little difference between natural and hatchery populations.

Recommendation: Continue to collect as many carcasses as possible for the most accurate age composition data. Continue to assist hatchery staff with picking eyed eggs to obtain fecundity estimates for each spawned female. Collect other biological data (length, run timing, spawn timing, DNA samples, juvenile parr production, smolt trapping, and life stage survival) to continue the documentation of effects (positive or negative) that the hatchery program may have on the natural population.
2. Documenting the success of hatchery origin fish spawning in the river has become an increasingly frequent topic among managers within the Snake River Basin and with NOAA Fisheries. Little, if any, data exists on this subject. With the hatchery population in the Tucannon River slowly replacing the natural population, we are offered an opportunity to study the effects of the hatchery spawners in the natural environment.

Recommendation: Participate in a reproductive success study for spring chinook being developed jointly by NOAA Fisheries/WDFW personnel. Continue to use snorkel surveys during the summer months to estimate spring chinook parr production in the river. Examine the relationship between redd counts and the following years parr production, smolt numbers and returning adults in context of the proportion of hatchery spawners in the river and publish the results.
3. Smolt trapping is our most valuable tool in estimating the number of fish emigrating from the river. Having accurate emigration estimates and knowing the confidence limits of those estimates is pertinent in calculating reliable survival rates.

Recommendation: Work with WDFW statisticians to refine our smolt trapping methods and statistical analyses. Publish statistical methods relating abiotic factors to smolt trap efficiency rates in the scientific literature.
4. Subbasin and recovery planning for listed species in the Tucannon River will identify factors limiting the spring chinook population and strategies to recover the population.
Development of a recovery goal for the population would be helpful to develop and evaluate strategies for habitat, hydropower, harvest and hatcheries.

Recommendation: Assist subbasin planning in the development of a recovery goal for spring chinook in the Tucannon River.
5. Smolt and adult detection capabilities for PIT tagged salmon within the Columbia and Snake River basins is becoming more widespread. These capabilities can help estimate survival rates for release groups to aid in evaluation of program success.

Recommendation: Utilize the SURPH2 PIT tag model software and present summaries of juvenile rates in future reports. Increase sample size of PIT tags if necessary, and document stray Tucannon fish above lower Granite Dam.
6. We have documented that hatchery juvenile (egg-parr-smolt) survival rates are considerably higher than naturally reared salmon, and hatchery smolt-to-adult return rates are much lower. We need to identify and address the factors that limit hatchery SAR's in order to meet mitigation goals.

Recommendation: Conduct a literature search and initiate a meeting between biologists working with spring chinook, both within and outside of the Snake River Basin, to compare survival rates from different watersheds under different rearing and release strategies.
Provide recommendations to improve SAR, or a list of recommended research topics for managers to consider that would provide answers to improve hatchery survival.

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## Appendix A

## Spring chinook captured, collected, or passed upstream at the Tucannon Hatchery trap in 2002

| Appendix A. Spring chinook salmon captured, collected, or passed upstream at the Tucannon Hatchery trap in 2002. (Trapping began in February; last day of trapping was September 30). |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Captured in trap |  | Collected for broodstock |  | Passed upstream |  |
| Date | Natural | Hatchery | Natural | Hatchery | Natural | Hatchery |
| 5/04 |  | 1 |  |  |  | 1 |
| 5/09 |  | 1 |  |  |  | 1 |
| 5/16 |  | 1 |  |  |  | 1 |
| 5/17 | 1 | 1 | 1 | 1 |  |  |
| 5/18 | 1 | 4 |  |  | 1 | 4 |
| 5/19 | 2 | 1 | 2 |  | 2 | 1 |
| 5/20 | 2 | 10 |  | 6 |  | 4 |
| 5/21 | 3 | 21 | 2 | 10 | 1 | 11 |
| 5/22 | 3 | 25 |  |  | 3 | 25 |
| 5/23 |  | 5 |  |  |  | 5 |
| 5/24 | 1 | 19 |  | 10 | 1 | 9 |
| 5/25 | 1 | 11 |  |  | 1 | 11 |
| 5/26 | 3 | 22 |  |  | 3 | 22 |
| 5/27 | 3 | 29 |  |  | 3 | 29 |
| 5/28 | 13 | 45 | 7 |  | 6 | 45 |
| 5/29 | 18 | 58 |  |  | 18 | 58 |
| 5/30 | 15 | 35 | 5 | 4 | 10 | 31 |
| 5/31 | 4 | 17 | 3 | 1 | 1 | 16 |
| 6/01 | 5 | 25 |  |  | 5 | 25 |
| 6/02 |  | 1 |  |  |  | 1 |
| 6/03 | 5 | 20 | 4 | 3 | , | 17 |
| 6/04 | 7 | 27 |  |  | 7 | 27 |
| 6/05 | 2 | 8 | 2 | 1 |  | 7 |
| 6/06 | 9 | 37 |  |  | 9 | 37 |
| 6/07 | 3 | 13 | 3 | 3 |  | 10 |
| 6/08 | 1 | 3 |  |  | 1 | 3 |
| 6/10 |  | 2 |  | 2 |  |  |
| 6/11 |  | 7 |  |  |  | 7 |
| 6/12 | 1 | 20 | 1 | 1 |  | 19 |
| 6/13 | 8 | 17 | 5 |  | 3 | 17 |


| 6/14 | 6 | 11 | 5 |  | 1 | 11 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 6/15 | 7 | 18 |  |  | 7 | 18 |
| 6/16 | 5 | 5 |  |  | 5 | 5 |
| 6/17 |  | 9 |  |  |  | 9 |
| 6/18 | 7 | 3 | 3 | 2 | 4 | 1 |
| 6/19 |  | 1 |  |  |  | 1 |
| 6/21 | 5 | 7 | 4 | 1 | 1 | 6 |
| 6/22 | 3 | 7 |  |  | 3 | 7 |
| 6/23 | 2 | 4 |  |  | 2 | 4 |
| 6/24 | 2 | 4 | 2 |  |  | 4 |
| 6/25 |  | 4 |  |  |  | 4 |
| 6/26 | 1 |  |  |  | 1 |  |
| 6/27 | 1 | 3 |  |  | 1 | 3 |
| 6/28 | 2 | 3 |  |  | 2 | 3 |
| 7/01 |  | 2 |  |  |  | 2 |
| 7/03 |  | 1 |  | 1 |  |  |
| 7/05 |  | 1 |  | 1 |  |  |
| 7/09 | 1 | 1 | 1 | 1 |  |  |
| 7/15 | 3 | 3 | 2 | 2 | 1 | 1 |
| 7/29 | 1 | 1 |  | 1 | 1 |  |
| 7/31 | 1 |  |  |  | 1 |  |
| Appendix A (continued). Spring chinook salmon captured, collected, or passed upstream at the Tucannon Hatchery trap in2002. |  |  |  |  |  |  |
|  | Captured in trap |  | Collected for broodstock |  | Passed upstream |  |
| Date | Natural | Hatchery | Natural | Hatchery | Natural | Hatchery |
| 8/11 |  | 3 |  |  |  | 3 |
| 8/15 |  | 1 |  |  |  | 1 |
| 8/22 | 3 |  |  |  | 3 |  |
| 8/27 | 4 | 1 |  |  | 4 | 1 |
| 8/28 | 2 | 1 |  |  | 2 | 1 |
| 8/29 | 4 | 1 |  |  | 4 | 1 |
| 8/30 | 1 | 1 |  |  | 1 | 1 |
| 9/04 |  | 1 |  | 1 |  |  |
| 9/07 |  | 5 |  |  |  | 5 |
| 9/09 |  | 3 |  | 1 |  | 2 |
| 9/10 | 2 |  |  |  | 2 |  |
| 9/12 | 3 | 5 |  |  | 3 | 5 |
| 9/13 |  | 1 |  |  |  | 1 |
| 9/15 | 1 | 1 |  |  | 1 | 1 |
| 9/16 |  | 2 |  | 2 |  |  |
| Totals | 178 | 600 | 52 | 55 | 126 | 545 |
| Corrected \#'s after spawning ${ }^{1}$ | 168 | 610 | 42 | 65 | 126 | 545 |

${ }^{1}$ Ten fish with intact adipose fins were subsequently identified to be of hatchery origin based on scale pattern analysis.

## Appendix B

## Movements of ten radio tagged female captive brood adults released into the Tucannon River, 2002

| Appendix B. Movements of ten radio tagged female captive brood adults released into the Tucannon River, 2002. |  |  |  |
| :---: | :---: | :---: | :---: |
| Channel/Code Date | Tucannon Rkm | Location | Comments |
| 9/165 $7 / 16 / 02$ $8 / 20 / 02$ $8 / 27 / 02$ $8 / 30 / 02$ $9 / 05 / 02$ $9 / 09 / 02$ $9 / 13 / 02$ $9 / 16 / 02$ $9 / 20 / 02$ $9 / 23 / 02$ $9 / 25 / 02$ | $\begin{aligned} & 74.5 \\ & 72.0 \\ & 73.0 \\ & 73.0 \\ & 72.8 \\ & 72.9 \\ & 72.9 \\ & 72.9 \\ & 72.9 \\ & 72.9 \\ & \hline \end{aligned}$ | Lyons Ferry Hatchery Panjab Bridge $3^{\text {rd }}$ Cattle guard 100 m above C.C. Br. 100 m above C.C. Br. Below C.C. Bridge Below C.C. Bridge Below C.C. Bridge Below C.C. Bridge Below C.C. Bridge Below C.C. Bridge | Length at tagging - 58.0 cm . <br> Released into river. <br> Drive by. <br> Between campground and cattle guard. <br> R.B. lower end of habitat site, by new redd. <br> L.B. below rocks, with wild male. <br> Area where she was digging now small TD. <br> Fungused eyes, fins, tail frayed. <br> Recovered tag and fish, $100 \%$ spent. |
| $9 / 167$ $7 / 16 / 02$ $8 / 20 / 02$ $8 / 27 / 02$ $8 / 30 / 02$ $9 / 05 / 02$ $9 / 09 / 02$ $9 / 13 / 02$ | $\begin{aligned} & 74.5 \\ & 73.2 \\ & 73.3 \\ & 72.9 \\ & 73.0 \\ & 73.0 \\ & \hline \end{aligned}$ | Lyons Ferry Hatchery Panjab Bridge HMA5-S Side Channel Log jam below log weir Cow Camp Bridge 100 m above C.C. Br. 100 m above C.C. Br. | Length at tagging - 55.5 cm . <br> Released into river. <br> Drive by. <br> By redd 2-6, with other fish, wild male close by. Recovered tag and fish - did not spawn. |
| 9/171 $7 / 16 / 02$ $8 / 20 / 02$ $8 / 27 / 02$ $8 / 30 / 02$ $9 / 05 / 02$ $9 / 09 / 02$ | $\begin{aligned} & 74.5 \\ & 73.2 \\ & 73.2 \\ & 73.2 \\ & 73.4 \end{aligned}$ | Lyons Ferry Hatchery Panjab Bridge HMA5-S Side Channel HMA5-S Side Channel Between C.C. \& C.G. 9 Above C.C. Br. . 35 km | Length at tagging - 56.5 cm . Released into river. <br> Drive by. <br> Log jam near 9/04/02JD test dig. |


| $9 / 13 / 02$ | 74.5 | Below Panjab Ck. mouth | Went down to pool w/ 9/183 then upstream. |
| :---: | :---: | :--- | :--- |
| $9 / 16 / 02$ | 73.6 | C.G. 9 lower entrance | Drive by. |
| $9 / 20 / 02$ | 73.6 | S.C. at C.G. 9 | Near new redds in S.C., not actively digging. |
| $9 / 23 / 02$ | 73.4 | Log jam above rock sill | Recovered tag and fish. Fish partially eaten. |
| $9 / 179$ |  |  |  |
| $7 / 16 / 02$ |  | Lyons Ferry Hatchery | Length at tagging -55.5 cm. |
| $8 / 20 / 02$ | 74.5 | Panjab Bridge | Released into river. |
| 8/27/02 |  | Ladybug Flat? | Couldn’t locate - heard chirps near Ladybug. |
| $8 / 30 / 02$ |  | Not Found | Couldn't locate. |
| $9 / 05-09 / 02$ | 77.7 | Ladybug Flat | Run and pool under poplar, fish moving around. |
| $9 / 13-16 / 02$ | 77.7 | Ladybug Flat | Under alder, about 25 mpstream of path sign. |
| $9 / 20 / 02$ | 77.7 | Ladybug Flat | Recovered tag only. Tag found between rocks. |
|  |  |  | Probably poached. |


| Appendix B (continued). Movements of ten radio tagged female captive brood adults released into the Tucannon River, 2002. |  |  |  |
| :---: | :---: | :---: | :---: |
| Channel/Code Date | Tucannon Rkm | Location | Comments |
| 9/183 $7 / 16 / 02$ $8 / 20 / 02$ $8 / 27 / 02$ $8 / 30 / 02$ $9 / 05 / 02$ $9 / 09 / 02$ $9 / 13 / 02$ $9 / 16 / 02$ $9 / 20 / 02$ | $\begin{aligned} & 74.5 \\ & 74.4 \\ & 74.5 \\ & 74.5 \\ & 74.5 \\ & 74.5 \\ & 74.5 \\ & 74.5 \\ & \hline \end{aligned}$ | Lyons Ferry Hatchery <br> Panjab Bridge <br> Below Panjab Bridge <br> Panjab Bridge <br> Above Panjab Bridge <br> Above Panjab Bridge <br> Above Panjab Bridge <br> Above Panjab Bridge <br> Above Panjab Bridge | Length at tagging - 52.0 cm . <br> Released into river. <br> In $2^{\text {nd }}$ pool above bridge. <br> In $2^{\text {nd }}$ pool above bridge. <br> In $2^{\text {nd }}$ pool above bridge. <br> Drive by. <br> Recovered tag only on bank - probably poached. |
| $9 / 20 / 184$ $7 / 16 / 02$ $8 / 20 / 02$ $8 / 27 / 02$ $8 / 30 / 02$ $9 / 05 / 02$ $9 / 09 / 02$ $9 / 13 / 02$ $9 / 16 / 02$ $9 / 20 / 02$ $9 / 23 / 02$ | $\begin{aligned} & 74.5 \\ & 74.5 \\ & 74.6 \\ & 74.6 \\ & 72.9 \\ & 69.0 \\ & 69.0 \\ & 68.7 \\ & 68.7 \\ & \hline \end{aligned}$ | Lyons Ferry Hatchery <br> Panjab Bridge <br> Panjab Bridge <br> Wilderness C.G. 1 <br> Info. Sign below C.G. 1 <br> Below Cow Camp Bridge <br> Below Cattle Chute Area <br> Below Cattle Chute Area <br> Above Camp Wooten Cabins <br> HMA 15 - Above Cabins | Length at tagging - 51.0 cm . Released into river. <br> Below redd 3-7MH, saw fish. <br> Upper end of camping area. <br> Drive by. <br> Fish fungused - will not live long. <br> In log jam at lower end of side channel. Drive by. |
| $\begin{gathered} \hline 9 / 192 \\ 7 / 16 / 02 \\ 8 / 20 / 02 \\ 8 / 27 / 02 \\ 8 / 30 / 02 \\ 9 / 05 / 02 \\ \hline \end{gathered}$ | $\begin{aligned} & 74.5 \\ & 74.4 \\ & 74.7 \end{aligned}$ | Lyons Ferry Hatchery <br> Panjab Bridge <br> Ladybug? <br> Below Panjab Bridge <br> Wild C.G. 1 | Length at tagging - 50.0 cm . <br> Released into river. <br> Couldn't locate - heard chirps near Ladybug. <br> Saw fish in pool across from 2 week old redd. |


| 9/09/02 | 74.5 | 100 m below main info. sign | Wood cutting area sign. |
| :---: | :---: | :---: | :---: |
| 9/13/02 | 74.6 | Below C.G. 1 | Beside redd 4-4, not on redd though. |
| 9/16/02 | 73.7 | C.G. 9 | Drive by. |
| 9/20/02 | 73.6 | S.C. at C.G. 9 | Near new redds in S.C., not actively digging. |
| 9/23/02 | 73.6 | S.C. at C.G. 9 | Near redd 5-3 (9-18-02JD). |
| 9/27/02 | 73.6 | Below C.G. 9 | Recovered tag and fish - 100\% spawned. |
| 9/193 |  |  |  |
| 7/16/02 |  | Lyons Ferry Hatchery | Length at tagging - 51.0 cm . |
| 8/20/02 | 74.5 | Panjab Bridge | Released into river. |
| 8/27/02 | 72.0 | $3{ }^{\text {rd }}$ Cattle Guard |  |
| 8/30/02 | 73.0 | 100 m above C.C. Bridge |  |
| 9/05/02 | 73.5 | Lower end C.G. 9 | Couldn't pinpoint - tag may be out of fish. |
| 9/09/02 | 73.5 | Across from house, above | In run 10 m above National Forest Boundary. |
| 9/13/02 | 73.5 | C.C. | Tag in otter den. |
| 9/23/02 | 73.5 | Across from house, above | Tag in den. |
|  |  | Across from house, above |  |
|  |  | C.C. |  |

Appendix B (continued). Movements of ten radio tagged female captive brood adults released into the Tucannon River, 2002.

| Channel/Code <br> Date | Tucannon <br> Rkm | Location | Comments |
| :---: | :---: | :--- | :--- |
| $\mathbf{9 / 2 0 3}$ |  |  |  |
| $7 / 16 / 02$ | 74.5 | Lyons Ferry Hatchery | Panjab Bridge |
| $8 / 20 / 02$ | 74.5 | Panjab Bridge | Rength at tagging -49.0 cm. |
| $8 / 27 / 02$ |  | Not Found | Neased into river. |
| $8 / 30 / 02$ |  | Not Found | Lost contact. |
| $9 / 05 / 02$ |  | Lyons Ferry Hatchery | Lest contact. |
| $\mathbf{9 / 2 0 5}$ |  | Panjab Bridge | Released into river. |
| $7 / 16 / 02$ | 74.5 | Below Panjab Bridge |  |
| $8 / 20 / 02$ |  | Not Found |  |
| $8 / 27 / 02$ | 76.6 | 1 km below Ladybug Flat | Fish holding under spruce over river. |
| $8 / 30 / 02$ | 76.6 | 1 km below Ladybug Flat | 50 m downstream of road 025. |
| $9 / 05 / 02$ | 76.6 | 1 km below Ladybug Flat | Recovered tag only under brush on bank. |
| $9 / 09 / 02$ |  |  | Probably poached. |
| $9 / 13 / 02$ |  |  |  |
|  |  |  |  |

## Appendix C

## Estimated Total Run-Size of Tucannon River Spring Chinook Salmon (1985-2002)

| Appendix C. Total estimated run-size of spring chinook salmon to the Tucannon River, 1985-2002. |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Run <br> Year | Wild <br> Jacks | Wild <br> Adults | Total <br> Wild | Hatchery <br> Jacks | Hatchery <br> Adults | Total <br> Hatchery | Total <br> Run-Size |
| 1985 | 0 | 591 | 591 | 0 | 0 | 0 | 591 |
| 1986 | 6 | 630 | 636 | 0 | 0 | 0 | 636 |
| 1987 | 6 | 576 | 582 | 0 | 0 | 0 | 582 |
| 1988 | 19 | 391 | 410 | 19 | 0 | 19 | 429 |
| 1989 | 2 | 334 | 336 | 83 | 26 | 109 | 445 |
| 199 | 0 | 494 | 494 | 22 | 238 | 260 | 754 |
| 1991 | 3 | 257 | 260 | 99 | 169 | 268 | 528 |
| 1992 | 12 | 406 | 418 | 15 | 320 | 335 | 753 |
| 1993 | 8 | 309 | 317 | 6 | 266 | 272 | 589 |
| 1994 | 0 | 98 | 98 | 5 | 37 | 42 | 140 |
| 1995 | 2 | 19 | 21 | 11 | 22 | 33 | 54 |
| 1996 | 2 | 145 | 147 | 15 | 70 | 85 | 232 |
| 1997 | 0 | 134 | 134 | 3 | 151 | 154 | 288 |
| 1998 | 0 | 85 | 85 | 16 | 43 | 59 | 144 |
| 1999 | 0 | 3 | 3 | 60 | 182 | 242 | 245 |
| 2000 | 14 | 68 | 82 | 16 | 241 | 257 | 339 |
| 2001 | 9 | 709 | 718 | 111 | 183 | 294 | 1,012 |
| 2002 | 9 | 341 | 350 | 11 | 644 | 655 | 1,005 |

## Appendix D

## Numbers and density estimates (fish/100 m²) of juvenile salmon counted by snorkel surveys in the Tucannon River in 2002

| Appendix D. Numbers and density estimates of subyearling and yearling natural salmon, and yearling hatchery chinook counted by snorkel surveys in the Tucannon River, 2002. |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Stratum | Site | Date | Number of Salmon |  |  | Snorkeled <br> Area (m ${ }^{2}$ ) | Density (fish/100m ${ }^{2}$ ) |  |  |
|  |  |  | Natural |  | Hatchery$>1+$ |  | Natural |  | Hatchery$>1+$ |
|  |  |  | 0+ | > $1+$ |  |  | 0+ | > $1+$ |  |
| Marengo $\downarrow$ | TUC01 | 8/12 | 37 | 1 | 0 | 423 | 8.74 | 0.24 | 0.00 |
|  | 01A | 8/12 | 28 | 0 | 0 | 336 | 8.33 | 0.00 | 0.00 |
|  | TUC02 | 8/12 | 32 | 0 | 0 | 656 | 4.88 | 0.00 | 0.00 |
|  | 02A | 8/12 | 89 | 0 | 0 | 556 | 16.00 | 0.00 | 0.00 |
|  | TUC03 | 8/12 | 13 | 0 | 0 | 520 | 2.50 | 0.00 | 0.00 |
|  | 03A | 8/12 | 88 | 0 | 0 | 482 | 18.26 | 0.00 | 0.00 |
| Hartsock $\downarrow$ | TUC04 | 8/12 | 49 | 0 | 0 | 423 | 11.58 | 0.00 | 0.00 |
|  | 04A | 8/12 | 80 | 0 | 0 | 498 | 16.06 | 0.00 | 0.00 |
|  | TUCO5 | 8/12 | 72 | 0 | 0 | 574 | 12.54 | 0.00 | 0.00 |
|  | 05A | 8/12 | 102 | 0 | 0 | 404 | 25.25 | 0.00 | 0.00 |
|  | TUC06 | 8/12 | 52 | 0 | 0 | 620 | 8.39 | 0.00 | 0.00 |
|  | 06A | 8/12 | 28 | 0 | 0 | 574 | 4.88 | 0.00 | 0.00 |
|  | TUC07 | 8/12 | 111 | 1 | 0 | 995 | 11.16 | 0.10 | 0.00 |
|  | 07A | 8/12 | 137 | 0 | 0 | 773 | 17.72 | 0.00 | 0.00 |
|  | TUC08 | 8/12 | 153 | 0 | 0 | 484 | 31.61 | 0.00 | 0.00 |
|  | 08A | 8/12 | 25 | 0 | 0 | 570 | 4.39 | 0.00 | 0.00 |
|  | TUC09 | 8/12 | 92 | 1 | 0 | 575 | 16.00 | 0.17 | 0.00 |
|  | 09A | 8/12 | 55 | 0 | 0 | 567 | 9.70 | 0.00 | 0.00 |
|  | TUC10 | 8/12 | 63 | 1 | 0 | 381 | 16.54 | 0.26 | 0.00 |
|  | 010A | 8/12 | 50 | 0 | 0 | 483 | 10.35 | 0.00 | 0.00 |
| HMA$\downarrow$ | TUC11 | 8/12 | 61 | 0 | 0 | 753 | 8.10 | 0.00 | 0.00 |
|  | 011A | 8/12 | 83 | 0 | 0 | 583 | 14.24 | 0.00 | 0.00 |
|  | TUC13 | 8/12 | 120 | 0 | 0 | 631 | 19.02 | 0.00 | 0.00 |
|  | 13A | 8/12 | 68 | 0 | 0 | 638 | 10.66 | 0.00 | 0.00 |
|  | TUC14 | 8/12 | 187 | 1 | 0 | 552 | 33.88 | 0.18 | 0.00 |
|  | 14A | 8/12 | 92 | 1 | 0 | 575 | 16.00 | 0.17 | 0.00 |
|  | TUC16 | 8/12 | 110 | 2 | 0 | 415 | 26.51 | 0.48 | 0.00 |
|  | 16A | 8/12 | 53 | 0 | 0 | 553 | 9.58 | 0.00 | 0.00 |
|  | TUC17 | 8/12 | 137 | 2 | 0 | 615 | 22.28 | 0.33 | 0.00 |


| 17 A | $8 / 12$ | 34 | 0 | 0 | 755 | 4.50 | 0.00 | 0.00 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TUC19 | $8 / 13$ | 15 | 0 | 0 | 606 | 2.48 | 0.00 | 0.00 |
| 19 A | $8 / 13$ | 96 | 0 | 0 | 495 | 19.39 | 0.00 | 0.00 |
| TUC20 | $8 / 13$ | 66 | 0 | 0 | 544 | 12.13 | 0.00 | 0.00 |
| 20A | $8 / 13$ | 42 | 0 | 0 | 571 | 7.36 | 0.00 | 0.00 |

Appendix D. Numbers and density estimates of subyearling and yearling natural salmon, and yearling hatchery chinook counted by snorkel surveys in the Tucannon River, 2002.

| Stratum | Site | Date | Number of Salmon |  |  | Snorkeled <br> Area (m) | Density (fish/100m ${ }^{2}$ ) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Natural |  | Hatchery$>1+$ |  | Natural |  | Hatchery$>1+$ |
|  |  |  | 0+ | > $1+$ |  |  | 0+ | > $1+$ |  |
| HMA (cont.) $\downarrow$ | TUC21 | 8/13 | 70 | 3 | 0 | 648 | 10.80 | 0.46 | 0.00 |
|  | 21A | 8/13 | 36 | 0 | 0 | 656 | 5.49 | 0.00 | 0.00 |
|  | TUC22 | 8/13 | 57 | 6 | 0 | 512 | 11.13 | 1.17 | 0.00 |
|  | 22A | 8/13 | 41 | 0 | 0 | 486 | 8.44 | 0.00 | 0.00 |
|  | TUC23 | 8/13 | 30 | 0 | 0 | 584 | 5.14 | 0.00 | 0.00 |
|  | 23A | 8/13 | 29 | 2 | 0 | 599 | 4.84 | 0.33 | 0.00 |
| Wilderness $\downarrow$ | TUC24 | 8/13 | 76 | 5 | 0 | 599 | 12.69 | 0.83 | 0.00 |
|  | 24A | 8/13 | 32 | 4 | 0 | 387 | 8.27 | 1.03 | 0.00 |
|  | TUC25 | 8/13 | 15 | 1 | 0 | 365 | 4.11 | 0.27 | 0.00 |
|  | 25A | 8/13 | 84 | 2 | 0 | 290 | 28.97 | 0.69 | 0.00 |
|  | TUC26 | 8/13 | 10 | 0 | 0 | 341 | 2.93 | 0.00 | 0.00 |
|  | 26A | 8/13 | 17 | 2 | 0 | 481 | 3.53 | 0.42 | 0.00 |
|  | TUC27 | 8/13 | 0 | 0 | 0 | 390 | 0.00 | 0.00 | 0.00 |
|  | 27A | 8/13 | 0 | 0 | 0 | 366 | 0.00 | 0.00 | 0.00 |
|  | TUC28 | 8/13 | 0 | 0 | 0 | 287 | 0.00 | 0.00 | 0.00 |
|  | 28A | 8/13 | 0 | 0 | 0 | 649 | 0.00 | 0.00 | 0.00 |
| Totals |  |  | 3,017 | 35 | 0 | 26,820 |  |  |  |

## Appendix E

## Recoveries of coded-wire tagged salmon released into the Tucannon River for the 1985-1997 brood years

| Appendix E. Observed and estimated recoveries of coded-wire tagged salmon released into the Tucannon River with percent return to the Tucannon Basin, out-of-basin returns, and estimated survival and exploitation rates for the 1985-1997 brood years. (Data from RMIS database.) |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Brood Year Smolts Released Fish/Lb CWT Codes ${ }^{1}$ Release Year | $\begin{gathered} \hline 1985 \\ 12,922 \\ 6.0 \\ 34 / 42 \\ 1987 \\ \hline \end{gathered}$ |  | 1986147,03710.0$33 / 25,41 / 46,41 / 48$1988 |  | $\begin{gathered} \hline 1987 \\ 151,100 \\ 9.0 \\ 49 / 50 \\ 1989 \\ \hline \end{gathered}$ |  |
| $\begin{array}{l}\text { Agency } \\ \text { (fishery/location) }\end{array}$ | Observed Number | Estimated Number | Observed Number | Estimated Number | Observed Number | Estimated Number |
| WDFW <br> Tucannon River <br> Kalama R., Wind R. <br> Fish Trap - F.W. <br> Treaty Troll <br> Lyons Ferry Hatch. ${ }^{2}$ <br> F.W. Sport <br> ODFW <br> Test Net, Zone 4 <br> Treaty Ceremonial <br> Three Mile, Umatilla R. <br> Spawning Ground <br> Fish Trap - F.W. <br> F.W. Sport <br> Hatchery <br> CDFO <br> Non-treaty Ocean Troll Mixed Net \& Seine Ocean Sport <br> USFWS <br> Warm Springs Hatchery Dworshak NFH | $32$ | 60 <br> 1 | 30 <br> 1 <br> 136 <br> 1 <br> 1 2 <br> 1 | $\begin{gathered} 2 \\ 287 \\ 4 \end{gathered}$ <br> 1 4 <br> 4 | 28 53 | 160 <br> 71 <br> 2 |
| Total Returns | 33 | 61 | 172 | 323 | 82 | 233 |
| Tucannon (\%) <br> Out-of-Basin (\%) <br> Commercial Harvest (\%) <br> Sport Harvest (\%) <br> Survival |  |  |  |  |  |  |

${ }^{1}$ WDFW agency code prefix is 63 .
${ }^{2}$ Fish trapped at TFH and held at LFH for spawning.

${ }^{1}$ WDFW agency code prefix is 63 .
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| Appendix E. Observed and estimated recoveries of coded-wire tagged salmon released into the Tucannon River with percent return to the Tucannon Basin, out-of-basin returns, and estimated survival and exploitation rates for the 1985-1997 brood years. (Data from RMIS database.) |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Brood Year <br> Smolts Released <br> Fish/Lb <br> CWT Codes ${ }^{1}$ <br> Release Year |  |  | $\begin{array}{r} 5 \\ 48 / 23, \end{array}$ | 9 $4,48 / 56$ | 199279,15114.0$48 / 10,48 / 55,49 / 05$1994 |  |
| Agency <br> (fishery/location) | Observed Number | Estimated Number | 1993  <br> Observed Estimated <br> Number Number |  | Observed Number | Estimated Number |
| WDFW <br> Tucannon River Kalama R., Wind R. Fish Trap - F.W. Treaty Troll Lyons Ferry Hatch. ${ }^{2}$ F.W. Sport | 24 | 24 | 2 | 2 | 11 45 | 34 49 |
| ODFW <br> Test Net, Zone 4 Treaty Ceremonial Three Mile, Umatilla R. Spawning Ground Fish Trap - F.W. F.W. Sport Hatchery | 1 | 3 3 | 1 | 1 | $\begin{aligned} & 2 \\ & 5 \\ & 2 \end{aligned}$ | $\begin{aligned} & 1 \\ & 4 \\ & 9 \\ & 2 \end{aligned}$ |
| CDFO <br> Non-treaty Ocean Troll Mixed Net \& Seine Ocean Sport |  |  | 1 | 2 |  |  |
| USFWS <br> Warm Springs Hatchery Dworshak NFH |  |  |  |  | 3 | 3 |
| Total Returns | 26 | 30 | 4 | 5 | 69 | 102 |
| Tucannon (\%) <br> Out-of-Basin (\%) <br> Commercial Harvest (\%) <br> Sport Harvest (\%) <br> Survival |  |  |  |  |  |  |

[^0]

[^1]| Appendix E. Observed and estimated recoveries of coded-wire tagged salmon released into the Tucannon River with percent return to the Tucannon Basin, out-of-basin returns, and estimated survival and exploitation rates for the 1985-1997 brood years. (Data from RMIS database.) |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Brood Year Smolts Released Fish/Lb CWT Codes ${ }^{1}$ Release Year |  | 8 $1 / 24-25$ | $\begin{gathered} \hline 1997 \\ 23,509 \\ 16.0 \\ 61 / 32 \\ 1999 \\ \hline \end{gathered}$ |  | $\begin{gathered} 1998 \\ 124,093 \\ 13.0 \\ 12 / 11 \\ 2000 \\ \hline \end{gathered}$ |  |
| Agency (fishery/location) | Observed Number | Estimated Number | Observed <br> Number | Estimated Number | Observed Number | $\begin{gathered} \hline \text { Estimated } \\ \text { Number } \\ \hline \end{gathered}$ |
| WDFW <br> Tucannon River <br> Kalama R., Wind R. <br> Fish Trap - F.W. <br> Treaty Troll <br> Lyons Ferry Hatch. ${ }^{2}$ <br> F.W. Sport <br> Non-treaty Ocean Troll <br> ODFW <br> Test Net, Zone 4 <br> Treaty Ceremonial <br> Three Mile, Umatilla R. <br> Spawning Ground <br> Fish Trap - F.W. <br> F.W. Sport <br> Hatchery <br> Columbia R. Gillnet <br> Columbia R. Sport <br> CDFO <br> Non-treaty Ocean Troll <br> Mixed Net \& Seine <br> Ocean Sport <br> USFWS <br> Warm Springs Hatchery Dworshak NFH | 43 <br> 96 <br> 1 <br> 2 | 139 <br> 99 <br> 1 <br> 2 |  | 85 <br> 46 <br> 2 <br> 1 <br> 8 9 | 30 <br> 7 <br> 1 <br> 6 <br> 1 | 135 <br> 7 <br> 2 $\begin{gathered} 6 \\ 35 \end{gathered}$ |
| Total Returns | 142 | 241 | 67 | 151 | 45 | 185 |
| Tucannon (\%) <br> Out-of-Basin (\%) <br> Commercial Harvest (\%) <br> Sport Harvest (\%) <br> Survival |  |  | $\begin{gathered} \hline 86.8 \\ 2.0 \\ 5.3 \\ 6.0 \\ 0.64 \\ \hline \end{gathered}$ |  | Incomplete Returns |  |

${ }^{1}$ WDFW agency code prefix is 63 .
${ }^{2}$ Fish trapped at TFH and held at LFH for spawning.

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U.S. Fish and Wildlife Service

Office of External Programs
4040 N. Fairfax Drive, Suite 130
Arlington, VA 22203


[^0]:    ${ }^{1}$ WDFW agency code prefix is 63 .
    ${ }^{2}$ Fish trapped at TFH and held at LFH for spawning.

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    ${ }^{2}$ Fish trapped at TFH and held at LFH for spawning.

