# Evaluation of Juvenile Salmon Production in 2018 from the Cedar River and Bear Creek 


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## Introduction

This report describes the emigration of five salmonid species from two tributaries in the Lake Washington watershed. The Cedar River tributary flows into the southern end of Lake Washington and Bear Creek flows into the Sammamish River, which in turn flows into the north end of Lake Washington (Figure 1). In each watershed, the abundance of juvenile migrants is the measure of freshwater production upstream from the trapping location.

In 1992, the Washington Department of Fish and Wildlife (WDFW) initiated an evaluation of sockeye fry migrants in the Cedar River to investigate the sources of low adult sockeye returns. In 1999, the Cedar River juvenile monitoring study was expanded in scope in order to include juvenile migrant Chinook salmon. This new scope extended the trapping season to a six-month period and, as a consequence, also allowed estimation of coho abundance and assessment of steelhead and cutthroat trout movement.

In 1997, WDFW initiated an evaluation of sockeye fry migrants in the Sammamish watershed. In 1997 and 1998, a juvenile trap operated in the Sammamish River during the downstream sockeye migration. In 1999, the monitoring site was relocated to Bear Creek to evaluate Chinook and sockeye production. Since 1999, the Bear Creek juvenile monitoring study estimates coho production and movement of steelhead and cutthroat trout.


Figure 1. Map of Lake Washington trap sites used to monitor abundance of juvenile migrant salmonids in the Cedar River and Bear Creek, near Renton and Redmond, respectively.

The primary goal of this study was to estimate the abundance of natural-origin sockeye fry, natural-origin Chinook, and natural-origin coho migrating from the Cedar River and Bear Creek into Lake Washington in 2018. These data allow an estimate of egg to fry survival of the 2017 brood. Daily abundance estimates also characterized the migration timing of each species into Lake Washington.

## Methods

## Fish Collection

## Trapping Gear and Operation

## Cedar River

A rotary screw trap operated in the lower Cedar River during the late winter and spring out migration period to assess migration of sockeye and Chinook fry, larger sub-yearling Chinook, coho, steelhead, and resident cutthroat. The Cedar River screw trap is a 5-ft diameter rotary scrap trap supported by a $12-\mathrm{ft} \times 30-\mathrm{ft}$ steel pontoon barge (Seiler et al., 2003). The screw trap operated at river mile (R.M) 1.6, under the I-405 Bridge (Figure 1) continuously for the entire migration period from mid-January through mid-July. The trap did not fish during 14 daylight and 17 night periods due to high river flows or public safety concerns. The trap also did not operate on 10 day and night periods to avoid catch of hatchery released sockeye salmon, which are extremely abundant and compromise our ability to count natural-origin fry when present. Debris jams stopped the trap from fishing during five daylight and 14 night periods. Catches were enumerated by species at dusk and dawn in order to discern diel movements. Fork lengths were randomly sampled on a weekly basis from all sockeye, Chinook, coho, and cutthroat.

Over the duration of the Cedar River juvenile monitoring study, trapping operations moved in response to changes in channel morphology. From 1992 to 2016, a small floating inclined-plane trap operated nightly from January through early April (Seiler et al., 2003). In the summer of 1998, dredging in the lower Cedar River forced the inclined-plane trap location to relocate in 1999 from R.M 0.25 to R.M 0.8 in order to operate under suitable river velocities. Beginning in 1999, WDFW also began operating a rotary screw trap at R.M 1.6 for the period April to July to enumerate Chinook salmon.

In contrast to previous years, during 2017 and 2018, we operated only a single rotary screw for the duration of the season at R.M 1.6. We made this change for three reasons. First, dredging in 2016 resulted in major channel modifications in the lower Cedar River that compromised the inclined-plane trapping site. Second, for the purposes of data comparability, we sought to use a single gear type over the course of the trapping season rather than one gear type early (incline plane) and different gear type late (rotary screw). Finally, the rotary screw trap simplifies trap staffing because unlike the incline plane trap, it does not require a trap operator to be present during all hours of operation. Thus, the inclined-plane trap was retired.

The Cedar River Hatchery at Landsburg releases sockeye fry into the Cedar River during the winter and spring to contribute to sockeye returns to the Cedar River and to help promote Lake Washington fisheries. The hatchery released 6.95 million sockeye fry into the Cedar River over 9 nights throughout the 2018 migration period. Hatchery staff released fry at three separate locations and often at two locations on the same night. In total, seven releases occurred at the lower location (R.M. 2.1), six from the middle location (R.M. 13.5), and five releases at upper location (R.M. 21.8). To avoid complications estimating hatchery and natural-origin components, the trap did not operate on hatchery release nights. We estimated missed catch of natural-origin sockeye during
hatchery nights when the trapped was pulled out of the water. Residual hatchery sockeye can migrate for up to three nights after a hatchery release (Kiyohara, 2013). Since hatchery sockeye fry are not externally identifiable as hatchery fish, we are unable to assess the rate of contribution of hatchery fry to natural origin catch and abundance. In some instances, larger fry size is indication of when fry are fed before their release. True abundance and survival of natural origin sockeye is likely lower than reported if hatchery fish delay their migration by one or more nights after a release.

## Bear Creek

A rotary screw trap operated from January 25 to July 10, 2018, ~100 yards downstream of the Redmond Way Bridge at the railroad trestle (Figure 1). The trap fished continuously, except for 14 periods when high flows and debris stopped the trap, and for six night or day periods, when the trap was pulled due to an extreme abundance of peamouth chub (Mylocheilus caurinus) during their spawning aggregations. Technicians enumerated the catch by species daily at dawn and dusk. Fork lengths were randomly sampled on a weekly basis from all Chinook, coho, and cutthroat.

Similar to the Cedar River, trapping operations changed in response to flow conditions, project objectives, and safety concerns. From January to April, an inclined-plane trap operated 100 yards downstream of the Redmond Way Bridge in years 1999 through 2011. A rotary screw trap fished for the remainder of the season from April to July. The inclined-plane trap was retired after 2011. The rotary screw trap now operates in late January to cover the early fry migration period as well as the spring parr and smolt migrations.

## PIT Tagging

During screw trap operation at both sites, a portion of natural-origin Chinook migrants and Steelhead smolts received passive integrated transponder (PIT) tags. Tagging occurred three to five times a week, between late April and June, and following standard protocols outlined by the Columbia River basin by the PIT Tag Steering Committee (2014). Chinook longer than 65 mm that displayed good physical health received a PIT tag. Fish were released the same day of capture or held overnight in perforated buckets suspended off the stern of the trap. Hatchery Chinook subyearlings were also PIT tagged at Issaquah Hatchery. 1,000 fish per week were released into Issaquah Creek covering three weeks: May 1, 8, and 24. All fish were tagged and held at the Issaquah hatchery one to two weeks in advance of their release.

The Hiram Chittenden Locks demarcate the freshwater to marine boundary between the Lake Washington watershed and Puget Sound (Figure 1). The locks have several PIT tag detection antennas in four smolt flumes and the adult fish ladder. One of two filling culverts in the large lock received a PIT tag antenna array in the fall of 2015. The following analysis includes detections from the filling culvert array. Here, the median migration date is the median date of all detected fish at all detection locations. The average travel time is the difference between release date and detection date. The survival rate is the total unique detections relative to the total released at each site.

## Trap Efficiencies

Throughout the season, mark and recapture of sockeye fry, Chinook, and coho provide an estimate of trap efficiency. Fry were marked in a solution of Bismarck brown dye ( 14 ppm for 1 hour) in an aerated bucket of stream water. Only healthy, marked fry were released above the trapping site while deceased or injured fish were removed. The trap efficiency for a day or night period is the total recaptured fish relative to the total number of released fish. In the Cedar River, efficiency trials were occasionally supplemented with hatchery sockeye fry to increase the size of release groups.

Larger Chinook parr were PIT tagged while coho were marked with alternating upper and lower, vertical and horizontal partial-caudal fin clips. A dilute solution of MS-222 and stream water anesthetized parr before clipping. Marks alternated on weekly intervals or more frequently with significant changes in river discharge. Beginning April 24, Chinook parr larger than 65 mm FL received PIT tags. Similar to fin clips, PIT tags enable stratified releases and recaptures. Before releasing, clipped Coho and PIT tagged Chinook fish recovered from marking in perforated buckets suspended behind the trap in calm river water.

Trap efficiency trials occurred every other night, with frequency determined by the catch of each species. Releases of fry, smolt, and parr in the Cedar River occurred 2 km upstream of the trap at the Maplewood Roadside Park. Fry were released 100 meters upstream of the Bear River trap at the Redmond Way Bridge and smolts 700 m upstream at the Union Hill Bridge.

## Analysis

The abundance of juvenile migrant salmonids is estimated using a mark-recapture approach and a single trap design (Volkhardt et al. 2007). The analysis is stratified by time in order to account for heterogeneity in capture rates throughout the season. The general approach is to estimate (1) missed catch, (2) efficiency strata, (3) abundance for each strata, (4) extrapolated migration prior to and post trapping, and (5) total production.

## Missed Catch

Total catch ( $\hat{u}_{i}$ ) during period $i$ is the actual catch ( $n$ ) summed with estimated missed catch ( $\hat{n})$ during trap outages. Missed catch is estimated using one of three different approaches depending on when a trap outage occurred because migration rates differ between the day and night: 1) entire missed night periods, 2) partial day or night periods, and 3) entire day periods when trap operations suspended.

## Missed Catch for Entire Night Periods

For night outages, missed catch is a straight-line interpolation between catches on adjacent nights. This approach assumes that abundance of migrating fish during the adjacent nights are similar to the outage period. When the outage occurred on a single night, variance of the estimated catch is the variance of the mean catch on adjacent nights (Equation 1). When the outages occur
on consecutive nights, then missed catch is estimated from one or both adjacent night catches (Equation 2).

## Equation 1

$$
\operatorname{Var}\left(\bar{n}_{i}\right)=\frac{\sum\left(n_{i}-\bar{n}_{i}\right)^{2}}{k(k-1)}
$$

## Equation 2

$$
\operatorname{Var}\left(\bar{n}_{i}\right)=\frac{\sum\left(\hat{n}_{i}-\bar{n}_{i}\right)^{2}}{k(k-1)}+\frac{\sum \operatorname{Var}\left(\hat{n}_{i}\right)}{k}
$$

where:
$k=$ number of sample nights used in the interpolation,
$n_{i}=$ actual night catch of unmarked fish used to estimate the un-fished interval,
$\bar{n}_{i}=$ interpolated night catch estimate (mean of adjacent night catches), and
$\hat{n}_{i}=$ missed night catch (estimated) of unmarked fish used to estimate the un-fished interval

When the night catch estimate interpolates between two or more consecutive nights, variance for each interpolated catch estimate is scaled by the coefficient of variation ( $C V$ ) of mean catch for adjacent night fishing periods by the interpolated catch estimates using:

## Equation 3

$$
\operatorname{Var}\left(\hat{n}_{i}\right)=\left[\hat{n}_{i}\left(\frac{\sqrt{\operatorname{Var}\left(\bar{n}_{i}\right)}}{\bar{n}_{i}}\right)^{2}\right]
$$

## Missed Catch for Partial Day and Night Periods

When the trap operated intermittently, missed catch during the un-fished interval ( $\hat{n}_{i}$ ) estimated by:

## Equation 4

$$
\hat{n}_{i}=T_{i} * \bar{R}
$$

where:

$$
\begin{aligned}
& T_{i}=\text { Hours during non-fishing period } i \\
& \bar{R}=\text { Mean catch rate (fish/hour) from adjacent fished periods }
\end{aligned}
$$

Variance associated with $\hat{n}_{i}$ estimated by:

## Equation 5

$$
\operatorname{Var}\left(\hat{n}_{i}\right)=T_{i}^{2} * \operatorname{Var}(\bar{R})
$$

Variance of the mean catch rate ( $\bar{R}$ ) for $k$ adjacent fishing periods is:
Equation 6
$\operatorname{Var}(\bar{R})=\frac{\sum_{i=1}^{i=k}\left(R_{i}-\bar{R}\right)^{2}}{k(k-1)}$

## Missed Catch for Entire Day Periods

Missed day-time catches in the inclined-plane trap were estimated by multiplying the previous night catch by the proportion of the 24 -hour catch caught during the day. This proportion $\left(\mathrm{F}_{\mathrm{d}}\right)$ estimated as:

## Equation 7

$$
\hat{F}_{d}=\frac{T_{d}}{\bar{Q}^{-1} T_{n}+T_{d}}
$$

Variance in the day-to-night catch ratio is:

## Equation 8

$$
\operatorname{Var}\left(\hat{F}_{d}\right)=\frac{\operatorname{Var}(\bar{Q}) T_{n}^{2} T_{d}^{2}}{\bar{Q}^{4}\left(\frac{1}{\bar{Q}} T_{n}+T_{d}\right)^{4}}
$$

where:
$T_{n}=$ hours of night during 24 hour period,
$T_{d}=$ hours of day during 24 hour period, and
$\bar{Q}_{d}=$ bi-weekly day-to-night catch ratio.

## Efficiency Strata

Stratification of capture and recapture data was necessary to accommodate for changes in trap efficiency over the season. These changes result from a number of factors including river flows, turbidity, and fish size. However, when using a mark-recapture approach to estimate abundance, precision of the estimate increases with the number of recaptures. A manufactured drawback of stratification can be a large variance associated with the estimate. Therefore, a $G$-test was used to determine whether to pool or hold separate adjacent efficiency trials (Sokal and Rohlf 1981).

Of the marked fish $(M)$ released in each efficiency trial, a portion are recaptured ( $m$ ) and a portion are not seen $(M-m)$. If the seen:unseen [ $m:(M-m)$ ] ratio differs between trials, the trial periods were considered as separate strata. However, if the ratio did not differ between trials, the two trials were pooled into a single stratum. A $G$-test determined whether adjacent efficiency trials were statistically different $(\alpha=0.05)$. Trials that did not differ were pooled and the pooled group
compared to the next adjacent efficiency trial. Trials that did differ were held separately. Pooling of time-adjacent efficiency trials continued iteratively until the seen:unseen ratio differed between time-adjacent trials. Once a significant difference was identified, the pooled trials were assigned to one strata and the significantly different trial indicated the beginning of the next strata.

## Abundance for Each Strata

The abundance of juvenile migrants for a given strata $h$ was calculated from maiden catch (actual and missed, $\hat{u}_{h}$ ), marked fish released in that strata ( $M_{h}$ ), and marked fish recaptured in that strata $\left(m_{h}\right)$. Abundance was estimated using a Bailey estimator appropriate for single trap designs (Carlson et al. 1998).

## Equation 9

$$
\hat{U}_{h}=\frac{\hat{u}_{h}\left(M_{h}+1\right)}{m_{h}+1}
$$

Variance associated with the Bailey estimator was modified to account for variance of the estimated catch during trap outages:

Equation 10

$$
V\left(\hat{U}_{h}\right)=V\left(\hat{u}_{h}\right)\left(\frac{\left(M_{h}+1\right)\left(M_{h} m_{h}+3 M_{h}+2\right)}{\left(m_{h}+1\right)^{2}\left(m_{i}+2\right)}\right)+\left(\frac{\left(M_{h}+1\right)\left(M_{h}-m_{h}\right) \hat{u}_{h}\left(\hat{u}_{h}+m_{h}+1\right)}{\left(m_{h}+1\right)^{2}\left(m_{h}+2\right)}\right)
$$

Maiden catch $\left(\hat{u}_{h}\right)$ was the sum of all actual and estimated catch during strata $h$. Variance of the catch [ $V\left(\hat{u}_{h}\right)$ ] was the sum of all estimated catch variances during strata $h$.

## Extrapolate Migration Prior to and Post Trapping

Modality of the trap catches suggested that migration outside the period of trap operation was minimal. Pre- and post-trapping migrations were estimated using linear extrapolation.

## Equation 11

$$
\hat{N}_{e}=\frac{\sum_{d=1}^{d=k} \hat{N}_{d}}{k} * \frac{t}{2}
$$

Variance of the extrapolation estimated as:

Equation 12

$$
V\left(\hat{N}_{e}\right)=\frac{\sum_{d=1}^{d=k}\left(\hat{N}_{d}-\bar{N}\right)^{2}}{k(k-1)} *\left(\frac{t}{2}\right)^{2}
$$

where:

$$
\begin{aligned}
\hat{N}_{d}= & \text { Daily migration estimates, } \\
k \quad= & \text { Number of daily migration estimates used in calculation, and } \\
t \quad= & \text { Number of days between assumed start/end of migration and the first/last } \\
& \text { day of trapping. }
\end{aligned}
$$

Pre- and post-season migration is based on the first and last five days of the catch. The assumed migration for sockeye was January 1 to June 30 on the Cedar River and January 1 to April 30 on Bear Creek. The assumed migration for Chinook in both watersheds was January 1 to July 13. Preand post-season migration was not estimated for coho or cutthroat.

## Total Production

Total production is the sum of the stratified abundance estimates for all $k$ strata and the extrapolated migration estimates. Confidence intervals and coefficient of variation associated with abundances are calculated from the variance:
Equation 13

$$
\hat{N}=\hat{N}_{\text {before }}+\sum_{h=1}^{h=k} \hat{U}_{h}+\hat{N}_{\text {after }}
$$

## Egg-to-Migrant Survival and Productivity

Egg-to-migrant survival is the abundance natural-origin juvenile migrants (age 0+) relative to the previous fall egg deposition by female adult spawners. The potential egg deposition (PED) is the product of the number of female spawners and their fecundity. Weekly fall spawning surveys estimate the number of sockeye spawners (assuming $50 \%$ are female) in Cedar River and Bear Creek. Cedar River sockeye fecundity during the broodstock collection for the hatchery was 3,050 eggs per female in 2017 (Sedgwick, 2018). The fecundity of Bear Creek sockeye are assumed to be the same as the fecundity of Cedar River broodstock sockeye.

Productivity for Chinook in both the Cedar River and Bear Creek is the number of age $0+$ out migrants produced per female spawner. Two life-history forms of sub-yearling Chinook salmon are observed in Puget Sound: small fry that migrate immediately after emergence and larger parr that spend several weeks to months rearing in freshwater streams. Fry are defined as fish emigrating between January and early April ( $8^{\text {th }}$ ) and larger parr are defined as fish emigrating between early April ( $9^{\text {th }}$ ) and July. Here, Chinook freshwater productivity is the number of migrants (both fry and parr combined) per female. The number of female Chinook is based on weekly fall redd counts and assumed to represent one female per redd. Average fecundity for the Cedar River and Bear Creek is assumed to be similar to the fecundity of Soos Creek Hatchery Chinook on the Green River (4,500 eggs per female). In recent years, the survival rate of Chinook appears high (e.g. $61.9 \%$ in 2011 Cedar). A possible source of overestimating the survival rate is underestimating fecundity. We measured fecundity in the Lake Washington basin at the Issaquah hatchery from 2014 to 2016 ( $\mathrm{N}=280$ females). Average fecundity during this period exceeded 4,500 eggs per female (Issaquah median $=5,222$; mean $=5,265$; standard deviation $=1,316$ ). Fecundity in each female typically varies as a function of body size and age. The relationship between female body size ( POH in mm ) and fecundity can be explained using a power function
(Fecundity $=0.0438 * \mathrm{POHmm}{ }^{1.8021}, \mathrm{R}^{2}=0.44$ ). For each year and stream, we estimated fecundity for each carcass on the spawning ground based on the POH length (carcass length data provided by A. Bosworth) and then calculated the average fecundity for the population based on the 2014 2016 measurements.

## Cedar River

## Sockeye

## Production Estimate

We estimated $8,725,471 \pm 1,434,498( \pm 95 \% \mathrm{CI})$ natural-origin sockeye fry entered Lake Washington from the Cedar River in 2018 (Table 1, Appendix A1). This estimate includes a small pre-season estimate of 16,500 fry between January 1 and January 12. Fry migration began prior to our first day of trapping as noted by sockeye catches on the first night (Figure 2). The majority of the fry migrated between March $1^{\text {st }}$ to the $23^{\text {rd }}$ (Figure 2). The median migration date for naturalorigin sockeye was March $11^{\text {th }}$. Cedar River sockeye fry are migrating earlier over the last 3 decades ( -9.6 days per decade, Table ). However, the 2018 run was 4 days later relative to the average run timing observed over the last 10 years (March $7^{\text {th }}$ ).

Efficiency data were aggregated into four release strata from twelve efficiency trials of sockeye fry (Table 1, Appendix A). Weekly recapture rates of the screw trap during the winter season appear consistent from year to year with the change in methods to a single gear type. Average efficiency rates in 2018 were $2.43 \%$. This rate closely matches efficiency rates observed in 2017 (2.82\%).

Table 1. Abundance of natural-origin sockeye fry entering Lake Washington from the Cedar River in 2018. Table includes; total catch (actual plus estimated), abundance of fry migrants, $95 \%$ confidence intervals (C.I.), coefficient of variation ( $C V$ ), and trap efficiency.

| Dates | Total catch | Abundance | Lower CI | Upper CI | $C V$ | $E f f$. |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Pre trapping Jan 1-12 | -- | 16,500 | 14,142 | 18,858 | $7.3 \%$ | -- |
| Jan 12- Feb 2 | 6,206 | 137,911 | 88,777 | 187,045 | $18.2 \%$ | $4.50 \%$ |
| Feb 3- Mar 23 | 175,591 | $7,153,257$ | $6,255,335$ | $8,051,179$ | $6.4 \%$ | $2.45 \%$ |
| Mar 24- Mar 30 | 20,709 | 516,112 | 366,679 | 665,545 | $14.8 \%$ | $4.01 \%$ |
| Mar 31- July 15 | 9,704 | 901,691 | 566,041 | $1,237,341$ | $19.0 \%$ | $1.08 \%$ |
| Total | 212,210 | $8,725,471$ | $7,290,974$ | $10,159,969$ | $5.68 \%$ | $2.43 \%$ |



Figure 2. Estimated daily migration of natural-origin sockeye fry migrating from the Cedar River into Lake Washington between January 1 and July 15, 2018. Pre-trapping migration estimates are included (Jan. 1-Jan. 12). Top panel displays changes in average fork length of natural-origin fry with vertical lines as $\pm 1$ standard deviation. Fork lengths of hatchery fry were taken at the holding ponds before their release. Middle panels show daily river flow and water temperature (or air) during this period (USGS gage \#12119000, air temperature in Renton, King County 31N).

The Cedar River Sockeye Hatchery released 6,947,607 million sockeye from February 18, 2018 through April 4, 2018 on nine nights (Table 2). Seven releases occurred at the lower location (River mile, R.M. 2.1), six from the middle location (R.M. 13.5) and five releases at the upper location (R.M. 21.8). The screw trap did not operate during release nights and part of the following day to reduce the impact on these fish and because their abundance can compromise our ability to accurately estimate natural-origin sockeye. Hatchery sockeye median migration date was March $14^{\text {th }}, 3$ days later than the median migration date of naturally produced sockeye in 2018 (Table 3). Hatchery fry were 3 to 5 mm longer in fork length (data from the hatchery) when they were released compared to natural origin fry captured at the trap (Figure 2).

Table 2. Release schedule of $6,947,607$ million hatchery sockeye from the Cedar River Sockeye Hatchery released at three different release points along the Cedar River in 2018: lower (river mile, R.M. 2.1), middle (R.M. 13.5) and upper location (R.M. 21.8)

| Release Date | Lower | Middle | Upper |
| :---: | ---: | :---: | ---: |
| 14-Feb | 130,326 | 485,977 | 242,328 |
| 27-Feb | 225,343 | 327,793 | 167,236 |
| 28-Feb | 720,163 |  | 248,132 |
| 12-Mar | 439,582 | 346,448 |  |
| 14-Mar | 413,410 | 730,934 | 275,299 |
| 19-Mar |  | 772,041 | 429,073 |
| 26-Mar | 694,580 |  |  |
| 28-Mar | 193,129 |  |  |
| 4-Apr |  | 105,813 |  |
|  | $2,816,533$ | $2,769,006$ | $1,362,068$ |

Table 3. Median migration dates of natural-origin, hatchery, and average combined sockeye fry from the Cedar River for trap years 1992 to 2018.

| Trap year | Natural | Hatchery | Combined | Diff (H-N) |
| :---: | :---: | :---: | :---: | :---: |
| 1992 | $03 / 18$ | $02 / 28$ | $03 / 12$ | 19 |
| 1993 | $03 / 27$ | $03 / 07$ | $03 / 25$ | 20 |
| 1994 | $03 / 29$ | $03 / 21$ | $03 / 26$ | 8 |
| 1995 | $04 / 05$ | $03 / 17$ | $03 / 29$ | 19 |
| 1996 | $04 / 07$ | $02 / 26$ | $02 / 28$ | 41 |
| 1997 | $04 / 07$ | $02 / 20$ | $03 / 16$ | 46 |
| 1998 | $03 / 11$ | $02 / 23$ | $03 / 06$ | 16 |
| 1999 | $03 / 30$ | $03 / 03$ | $03 / 15$ | 27 |
| 2000 | $03 / 27$ | $02 / 23$ | $03 / 20$ | 33 |
| 2001 | $03 / 10$ | $02 / 23$ | $03 / 08$ | 15 |
| 2002 | $03 / 25$ | $03 / 04$ | $03 / 19$ | 21 |
| 2003 | $03 / 08$ | $02 / 24$ | $03 / 03$ | 12 |
| 2004 | $03 / 21$ | $02 / 23$ | $03 / 15$ | 27 |
| 2005 | $03 / 02$ | $02 / 23$ | $03 / 01$ | 7 |
| 2006 | $03 / 20$ | $03 / 06$ | $03 / 16$ | 14 |
| 2007 | $03 / 23$ | $02 / 20$ | $02 / 26$ | 31 |
| 2008 | $03 / 16$ | $03 / 06$ | $03 / 15$ | 10 |
| 2009 | $03 / 19$ | $03 / 06$ | $03 / 13$ | 13 |
| 2010 | $03 / 07$ | $03 / 08$ | $03 / 07$ | -1 |
| 2011 | $03 / 25$ | $02 / 18$ | $03 / 01$ | 35 |
| 2012 | $03 / 22$ | $03 / 08$ | $03 / 18$ | 14 |
| 2013 | $03 / 07$ | $03 / 06$ | $03 / 07$ | 1 |
| 2014 | $03 / 02$ | $03 / 11$ | $03 / 04$ | -9 |
| 2015 | $03 / 07$ | $03 / 12$ | $03 / 07$ | -5 |
| 2016 | $03 / 07$ | $03 / 14$ | $03 / 14$ | -7 |
| 2017 | $02 / 28$ | $03 / 08$ | $03 / 03$ | -8 |
| 2018 | $03 / 11$ | $03 / 14$ | $03 / 13$ | -3 |
|  | 03 |  |  | 14 |

## Egg-to-Migrant Survival of Natural-Origin Fry

Egg-to-migrant survival of the 2017 brood Cedar River sockeye was $18.29 \%$ (Table 4). Survival was based on 8.725 million natural-origin fry from a potential 47.7 million eggs deposited by 15,645 females (A. Bosworth, Washington Department of Fish and Wildlife, personal communication). Average fecundity for the 2017 brood was 3,050 eggs per female sockeye (M. Sedgwick, Washington Department of Fish and Wildlife, personal communication).

Salmon eggs and alevins incubating within streambed redds are susceptible to flooding and scour, so peak winter discharges often explain annual variation in egg to fry survival. River flows surpassed known scouring thresholds ( $2,200 \mathrm{ft}^{3} \mathrm{sec}^{-1}$, Gendaszek et al. 2017) during egg incubation (November through February). Peak flows during December 2017 were at scour thresholds of $2,310 \mathrm{ft}^{3} \mathrm{sec}^{-1}$. Peak discharge on February $6^{\text {th }}$ reached $2330 \mathrm{ft}^{3} \mathrm{sec}^{-1}$ (Table 4). Trap technicians noticed fry with eggs attached (alevins) throughout February during higher flows. An unusual mortality event of 127 fry was observed on February $2^{\text {nd }}$ and the fry submitted to the WDFW pathology lab. All appeared to have died from coagulated yolk syndrome, suggesting a scour event killed them. The majority of the migration occurred when daily flows were more moderate (500$1,100 \mathrm{ft}^{3} \mathrm{~s}^{-1}$ ) in mid-March (Figure 2).

Table 4. Egg-to-migrant survival of natural-origin sockeye fry in the Cedar River and peak mean daily flows during egg incubation period for brood years 1991-2017. Incubation period is defined as November 1 to February 28. USGS monitors river flow continuously in Renton at station 12119000.

| Brood yr | Spawners | Females | Fecundity | Egg deposition | Fry | Survival | Peak flow | Flow date |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1991 | 76,592 | 38,296 | 3,282 | $125,687,226$ | $9,800,000$ | $7.80 \%$ | 2,060 | $1 / 28 / 1992$ |
| 1992 | 99,849 | 49,924 | 3,470 | $173,237,755$ | $27,100,000$ | $15.64 \%$ | 1,570 | $1 / 26 / 1993$ |
| 1993 | 74,677 | 37,338 | 3,094 | $115,524,700$ | $18,100,000$ | $15.67 \%$ | 927 | $1 / 14 / 1994$ |
| 1994 | 107,767 | 53,883 | 3,176 | $171,133,837$ | $8,700,000$ | $5.08 \%$ | 2,730 | $12 / 27 / 1994$ |
| 1995 | 21,443 | 10,721 | 3,466 | $37,160,483$ | 730,000 | $1.96 \%$ | 7,310 | $11 / 30 / 1995$ |
| 1996 | 228,391 | 114,196 | 3,298 | $376,616,759$ | $24,390,000$ | $6.48 \%$ | 2,830 | $1 / 2 / 1997$ |
| 1997 | 102,581 | 51,291 | 3,292 | $168,848,655$ | $25,350,000$ | $15.01 \%$ | 1,790 | $1 / 23 / 1998$ |
| 1998 | 48,385 | 24,193 | 3,176 | $76,835,676$ | $9,500,000$ | $12.36 \%$ | 2,720 | $1 / 1 / 1999$ |
| 1999 | 21,755 | 10,877 | 3,591 | $39,060,930$ | $8,058,909$ | $20.63 \%$ | 2,680 | $12 / 18 / 1999$ |
| 2000 | 146,060 | 73,030 | 3,451 | $252,025,754$ | $38,447,878$ | $15.26 \%$ | 627 | $1 / 5 / 2001$ |
| 2001 | 117,225 | 58,613 | 3,568 | $209,129,787$ | $31,673,029$ | $15.15 \%$ | 1,930 | $11 / 23 / 2001$ |
| 2002 | 192,395 | 96,197 | 3,395 | $326,590,484$ | $27,859,466$ | $8.53 \%$ | 1,410 | $2 / 4 / 2003$ |
| 2003 | 109,164 | 54,582 | 3,412 | $186,233,926$ | $38,686,899$ | $20.77 \%$ | 2,039 | $1 / 30 / 2004$ |
| 2004 | 114,839 | 57,419 | 3,276 | $188,106,200$ | $37,027,961$ | $19.68 \%$ | 1,900 | $1 / 18 / 2005$ |
| 2005 | 49,846 | 24,923 | 3,065 | $76,388,804$ | $10,861,369$ | $14.22 \%$ | 3,860 | $1 / 11 / 2006$ |
| 2006 | 105,055 | 52,527 | 2,910 | $152,854,370$ | $9,246,243$ | $6.05 \%$ | 5,411 | $11 / 9 / 2006$ |
| 2007 | 45,066 | 22,533 | 3,450 | $77,738,114$ | $25,072,141$ | $32.25 \%$ | 1,820 | $12 / 3 / 2007$ |
| 2008 | 17,300 | 8,650 | 3,135 | $27,118,177$ | $1,630,081$ | $6.01 \%$ | 9,390 | $1 / 8 / 2009$ |
| 2009 | 12,501 | 6,250 | 3,540 | $22,125,910$ | $12,519,260$ | $56.58 \%$ | 2,000 | $11 / 19 / 2009$ |
| 2010 | 59,795 | 29,898 | 3,075 | $91,935,489$ | $4,517,705$ | $4.91 \%$ | 5,960 | $1 / 18 / 2011$ |
| 2011 | 23,655 | 11,827 | 3,318 | $39,243,121$ | $14,763,509$ | $37.62 \%$ | 2,780 | $1 / 30 / 2012$ |
| 2012 | 88,974 | 44,487 | 3,515 | $156,371,805$ | $55,793,120$ | $35.68 \%$ | 1,513 | $12 / 7 / 2012$ |
| 2013 | 140,682 | 70,341 | 3,362 | $236,486,442$ | $37,975,769$ | $16.06 \%$ | 1,762 | $11 / 20 / 2013$ |
| 2014 | 10,450 | 5,225 | 3,368 | $17,597,800$ | $13,878,932$ | $78.87 \%$ | 2,162 | $1 / 8 / 2015$ |
| 2015 | 7,191 | 3,596 | 3,070 | $11,038,185$ | $2,163,843$ | $19.60 \%$ | 4,661 | $12 / 7 / 2015$ |
| 2016 | 7,573 | 3,787 | 3,144 | $11,904,756$ | $2,530,668$ | $21.26 \%$ | 2,140 | $2 / 10 / 2017$ |
| 2017 | 31,290 | 15,645 | 3,050 | $47,717,250$ | $8,725,471$ | $18.29 \%$ | 2,330 | $2 / 6 / 2018$ |

## Chinook

## Production Estimate

The total production of Chinook sub-yearling (parr and fry) in 2018 was 524,378 $\pm 78,450$ ( $\pm 95 \%$ C.I.). Here, Chinook fry are defined as those fish emigrating from January to April $8^{\text {th }}$ and Chinook parr are defined as fish emigrating after April $9^{\text {th }}$ (Figure 3). During the transition period in April, the overall migration decreases and larger size parr start to appear in the catch (Figure 3). The fry component was $94 \%$ of all natural Chinook production. Because trap efficiencies are not statistically different between sockeye and Chinook, sockeye releases were used to help estimate trap efficiency of Chinook fry (Kiyohara 2016). Trap efficiencies averaged $2.46 \%$ for sockeye fry.

An estimated 31,804 natural-origin Chinook parr passed the screw trap in 2018 (Table 5). This estimate is based on a total catch of 1,943 Chinook parr and trap efficiency of $6.11 \%$. Parr made up only $6 \%$ of the total sub-yearling migration. Chinook fry migration increased quickly over the season to one prominent peak in early March then slowly decreased for the remainder of season (Figure 3). Parr displayed sporadic movements in June that averaged 1,000 fish per day in late May and early June (Figure 3).


Figure 3. Bottom panel: Estimated daily migration of Chinook fry and parr from the Cedar River from January 1 to July 12, 2018. Pre-trapping migration estimates cover January 1-12. Parr life history type designation starts on April $9^{\text {th }}$ (plot arrow). Middle panels: Time series of mean daily water and air temperatures. River discharge from the USGS gaging station in Renton (Station 12119000). Top panel: Mean weekly Chinook body fork length with vertical lines as $\pm 1$ standard deviation and '. ' $\pm$ maximum and minimum weekly fork length.

Table 5. Abundance of Chinook migrants from Cedar River in 2018. Table includes total catch, abundance of fry and parr life history types, $95 \%$ confidence intervals (C.I.), coefficient of variation ( $C V$ ) and trap efficiencies (Eff.)

| Life history | Period | Catch | Abundance | Lower CI | Upper CI | $C V$ | Eff. |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Fry Jan 1-12 pre trapping | -- | 1,637 | 1,410 | 1,864 | $7.1 \%$ |  |  |
| Fry | Jan 13-Feb 2 | 443 | 9,844 | 6,484 | 13,204 | $17.4 \%$ | $4.50 \%$ |
| Fry | Feb 3- Mar 23 | 10,653 | 433,986 | 385,346 | 482,626 | $5.7 \%$ | $2.45 \%$ |
| Fry | Mar 24 -Mar 30 | 712 | 17,745 | 13,194 | 22,296 | $13.1 \%$ | $4.01 \%$ |
| Fry | Mar 31- April 8 | 316 | 29,362 | 18,117 | 40,607 | $19.5 \%$ | $1.08 \%$ |
| Parr | April 9-July 15 | 1,943 | 31,804 | 21,377 | 42,231 | $16.7 \%$ | $6.11 \%$ |
|  | Fry total: | 12,124 | 492,574 | 424,551 | 560,597 | $5.2 \%$ | $2.47 \%$ |
|  | Parr total: | 1,943 | 31,804 | 21,377 | 42,231 | $16.7 \%$ | $6.11 \%$ |
|  | Chinook total: | 14,067 | 524,378 | 445,928 | 602,828 | $5.0 \%$ | $2.68 \%$ |

## Size

Chinook migrant fork length (FL) ranged from 36 to 132 mm (Figure 3). Average fork length increased as water temperatures increased (Figure 3). Weekly average size of fry was 38 mm increasing to 42 mm FL by the first week in April (Figure 3). Chinook parr averaged 44 mm during the second week of April and increased to 96 mm by the first week in July.

## Productivity

Egg-to-migrant survival of the 2017 brood Cedar River Chinook was $14.2 \%$ (Table 6). Survival was based on 524,378 sub-yearling migrants and 3.69 million eggs from 819 female spawners (A. Bosworth, Washington Department of Fish and Wildlife, personal communication). The number of juvenile Chinook migrants produced per female (640) is near the $50^{\text {th }}$ percentile relative to the last 20 years (median $=571$ migrants per female, Table 6). The 2017 egg-to-migrant survival assuming 4,500 eggs per female is above 2025 goals (13.8\%) for the Cedar (WRIA 8 Conservation plan 2017). Egg-to-fry survival rates were encouragingly high during 2011-2015 brood years (survival rate range: $30-62 \%$ ), but the 2017 brood year appeared to have a lower survival ( $14.2 \%$ ) despite moderate flows. Cedar River discharge approached scouring thresholds ( $2,200 \mathrm{ft}^{3} \mathrm{sec}^{-1}$, Gendaszek et al. 2017) during egg incubation (November through February). Peak flows during November and December 2017 were 2,070 and $2,310 \mathrm{ft}^{3} \mathrm{sec}^{-1}$ (respectively). Peak discharge on February $6^{\text {th }}$ reached $2,330 \mathrm{ft}^{3} \mathrm{sec}^{-1}$ (Figure 3).

We calculated an alternative egg-to-migrant survival estimate using the relationship between body size and fecundity for adult females. From 2001-2017, we estimated average Chinook fecundity on the spawning grounds based on the post-orbital eye to hypural plate length $(\mathrm{POH})$ of female carcasses (Issaquah average $=5,265 ; \mathrm{SD}=1,316$, carcass length data at the spawning ground provided by A. Bosworth and A. David, Appendix A 4). This alternative formulation can produce a lower annual survival rate estimate than if we assume fecundity is 4,500 eggs per female. For instance, the Cedar River survival estimate in 2011 drops $16 \%$ from $61.9 \%$ to $45.9 \%$ under a POH based estimate of fecundity because larger sized females were found on the spawning ground (Table 6). For years when the survival rate is lower, the difference between the two calculations is
less noticeable ( $<1.5 \%$ ), but not negligible. The survival rate for the 2017 brood would be $12.0 \%$ instead of 14.2 \% and therefore, slightly below WRIA conservation plan 2025 goals.

Table 6. Abundance of Chinook fry and parr and productivity (juveniles per female) among brood years 1998 to 2017. Productivity is based on 4,500 eggs per females and weekly fall redd surveys. An alternative survival estimate uses Chinook fecundity on the spawning ground based on the post-orbital eye to hypural plate length $(\mathrm{POH})$ of female carcasses (data provided by A. Bosworth and A. David).

| Brood | Fry | Parr | Total | $\pm 95 \% \mathrm{CI}$ | \%Fry | \%Parr | Redds | Fry per <br> Female | Parr per Female | Total per <br> Female | Egg Survival | Alt. Egg Survival |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | 63,702 | 17,230 | 80,932 | 7,732 | 79\% | 21\% | 173 | 368 | 100 | 468 | 10.4\% |  |
| 1999 | 46,500 | 18,223 | 64,723 | 5,609 | $72 \%$ | 28\% | 182 | 255 | 100 | 356 | 7.9\% |  |
| 2000 | 10,833 | 21,416 | 32,249 | 5,220 | 34\% | 66\% | 53 | 204 | 404 | 608 | 13.5\% |  |
| 2001 | 79,799 | 39,875 | 119,674 | 41,349 | 67\% | 33\% | 398 | 201 | 100 | 301 | 6.7\% | 6.3\% |
| 2002 | 194,657 | 40,740 | 235,397 | 51,485 | 83\% | 17\% | 281 | 693 | 145 | 838 | 18.6\% | 14.8\% |
| 2003 | 65,752 | 55,124 | 120,876 | 2,518 | 54\% | 46\% | 337 | 195 | 164 | 359 | 8.0\% | 6.0\% |
| 2004 | 74,292 | 60,006 | 134,298 | 42,912 | 55\% | 45\% | 511 | 145 | 117 | 263 | 5.8\% | 4.4\% |
| 2005 | 98,967 | 18,592 | 117,559 | 16,233 | 84\% | 16\% | 339 | 292 | 55 | 347 | 7.7\% | 6.1\% |
| 2006 | 110,961 | 14,225 | 125,186 | 16,912 | 89\% | 11\% | 587 | 189 | 24 | 213 | 4.7\% | 3.7\% |
| 2007 | 705,583 | 64,208 | 769,791 | 76,106 | 92\% | 8\% | 899 | 785 | 71 | 785 | 19.0\% | 15.5\% |
| 2008 | 127,064 | 12,388 | 139,452 | 38,399 | 91\% | 9\% | 599 | 212 | 21 | 233 | 5.2\% | 3.8\% |
| 2009 | 115,474 | 36,916 | 152,390 | 13,058 | 76\% | 24\% | 285 | 405 | 130 | 535 | 11.9\% | 8.7\% |
| 2010 | 177,803 | 10,003 | 187,806 | 63,560 | 95\% | 5\% | 266 | 668 | 38 | 706 | 15.7\% | 11.0\% |
| 2011 | 863,595 | 38,919 | 902,514 | 165,973 | 96\% | 4\% | 324 | 2,665 | 120 | 2,786 | 61.9\% | 45.9\% |
| 2012 | 874,658 | 19,219 | 893,877 | 77,993 | 98\% | $2 \%$ | 433 | 2,020 | 44 | 2,064 | 45.9\% | 41.3\% |
| 2013 | 1,426,631 | 32,130 | 1,458,761 | 390,039 | 98\% | 2\% | 740 | 1,928 | 43 | 1,971 | 43.8\% | 33.1\% |
| 2014 | 326,901 | 20,762 | 347,663 | 90,223 | 94\% | 6\% | 232 | 1,409 | 89 | 1,499 | 33.3\% | 29.4\% |
| 2015 | 941,443 | 31,198 | 972,641 | 408,314 | 97\% | 3\% | 723 | 1,302 | 43 | 1,345 | 29.9\% | 23.6\% |
| 2016 | 151,262 | 23,457 | 174,719 | 37,722 | 87\% | 13\% | 418 | 362 | 56 | 418 | 9.3\% | 8.1\% |
| 2017 | 492,574 | 31,804 | 524,378 | 78,450 | 94\% | 6\% | 819 | 601 | 39 | 640 | 14.2\% | 12.0\% |

## Coho

## Production Estimate

Total coho age $1+$ smolt production was $179,946 \pm 52,442$ ( $\pm 95 \%$ C.I.) migrants (Table 7). The median migration date was May $7^{\text {th }}$, the peak migration day when we estimated more than 30,000 coho smolts. Production estimate for Coho was the most on record and is progressively increasing (Table 8). In comparison, the median run size between 2000 to 2017 was 60,621 smolts. Total catch (actual and estimated missed) of coho migrants in the screw trap was 5,848 . We observed two life history forms in the Cedar River: typical $1+$ yearling coho and sub-yearling coho fry and parr ( 62 total). Seven weeks of efficiency trials were aggregated into one strata with an efficiency of $3.25 \%$ (Table 7). This estimate includes only yearlings that moved past the screw trap (Figure 4). Fry and parr were not included in the estimate as their catch was very small.


Figure 4. Bottom panel: Estimated daily migration of yearling coho from the Cedar River in 2018 based on screw trap estimates from January 1 to July 12. Middle panels: Time series of mean daily air and water temperatures and river discharge from the USGS gaging station in Renton (Station 12119000). Top panel: Mean weekly Coho body fork length with vertical lines as $\pm 1$ standard deviation and '.' $\pm$ maximum and minimum weekly fork length. Age $1+$ smolts in filled points and age $0+$ fry and parr in open points.

Table 7. Abundance of coho smolt migrants from Cedar River in 2018. Table includes abundance of yearling migrants, $95 \%$ confidence intervals (C.I.), coefficient of variation (CV) and trap efficiency.

| Period | Total Catch | Abundance | Lower CI | Upper CI | $C V$ | Eff. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Jan 12 - July 15 | 5,848 | 179,946 | 127,504 | 232,388 | $14.87 \%$ | $3.25 \%$ |

## Size

Average fork length of all measured coho migrants smolts was 102 mm ; weekly averages ranged from 84 mm to 114 mm . Smolt migrants ranged from 76 mm to 149 mm FL (Figure 4). Sub-yearlings coho (fry and parr) ranged from 35 to 40 mm between February $2^{\text {nd }}$ and April $9^{\text {th }}$. Subyearling coho grew from an average of 40 to 74 mm between April $16^{\text {th }}$ and July $9^{\text {th }}$.

Table 8. Annual catch, abundance estimate, and $95 \%$ C.I. of natural-origin juvenile coho smolts emigrating from Cedar River from brood years 1997 to 2016.

| Brood | Tr | Tota | Start | End | Abundan |  | CI | CV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 1999 | 5,018 | 03/18 | 07/27 | 39,088 | 35,241 | 42,935 | 5.00\% |
| 1998 | 2000 | 2,446 | 04/27 | 07/13 | 32,169 | 30,506 | 33,833 |  |
| 1999 | 2001 | 6,262 | 04 | 07 | 82,462 | 60,293 | 104,661 | 13.70\% |
| 2000 | 2002 | 3,716 | 04/01 | 07/22 | 60,513 | 50,286 | 70,740 | 8.60\% |
| 2001 | 2003 | 3,964 | 04 | 07/12 | 74,507 | 58,947 | 90,067 | 10.70\% |
| 2002 | 2004 | 2,808 |  | 07 | 70,044 | 46,735 | 93,353 | 17.00\% |
| 2003 | 2005 | 2,918 | 04 | 07/28 | 72,643 | 42,725 | 102,561 | 21.40\% |
| 2004 | 2006 | 795 | 04/01 | 07/16 | 38,023 | 16,416 | 59,629 | 28.90\% |
| 2005 | 2007 | 482 | 04 | 07/20 | 33,994 | 8,291 | 59,697 | 40.80\% |
| 2006 | 2008 | 315 | 04 | 07/19 | 13,322 | 3,392 | 23,372 |  |
| 2007 | 2009 | 5,805 | 04/2 | 07/18 | 52,691 | 45,600 | 49,781 | 6.87\% |
| 2008 | 2010 | 6,528 | 04/22 | 07/04 | 83,060 | 70,049 | 96,071 | 7.99\% |
| 2009 | 2011 | 4,930 | 04/2 | 07/16 | 52,458 | 44,645 | 60,271 | 7.60\% |
| 2010 | 2012 | 2,912 | 04/1 | 07/14 | 48,168 | 38,493 | 57,843 | 10.25\% |
| 2011 | 2013 | 4,623 | 04/17 | 07/17 | 115,185 | 90,688 | 139,682 | 10.90\% |
| 2012 | 2014 | 8,071 | 04/1 | 07/16 | 129,666 | 104,393 | 154,940 | 9.94\% |
| 2013 | 2015 | 5,209 | 04/0 | 07/08 | 107,874 | 91,047 | 124,701 | 7.96\% |
| 2014 | 2016 | 2,720 | 04/14 | 07/14 | 60,621 | 41,862 | 79,379 | 15.79\% |
| 2015 | 2017 | 2,798 | 01/12 | 07/12 | 91,295 | 61,769 | 120,821 | 16.50\% |
| 2016 | 2018 | 5,848 | 01/12 | 07/15 | 179,946 | 127,504 | 232,388 | 14.87\% |

## Trout

Life history strategies used by trout in the Cedar River include anadromous, adfluvial, fluvial, and resident forms. Catches and estimates reported herein are for trout that were visually identified as either Oncorhynchus clarki (cutthroat trout) or Oncorhynchus mykiss (steelhead/rainbow trout). Steelhead smolts were identified when the fish had silver coloration upon capture. We did not identify trout fry to species or life-history type.

Six steelhead smolts, 120 juvenile cutthroat trout, 16 unidentifiable trout fry, and 4 adult cutthroat trout were captured in the screw trap. Catches were too few to estimate migrant abundance. Steelhead fork lengths ranged from 160 to 180 mm and averaged 166 mm . Juvenile cutthroat fork lengths ranged from 73 mm to 226 mm and averaged 151 mm .

## Incidental Catch

Twenty-three species of fish were documented in the Cedar River over the last 3 years. Other salmonids caught in the screw trap during 2018 include 259 hatchery Chinook parr and 16 pink fry (Oncorhynchus gorbuscha). Non-salmonid fishes encountered in the trap during 2018 include 47 lamprey (Lampetra spp.), 78 three-spine stickleback (Gasterosteus aculeatus), 183 sculpin (Cottus spp.), 7 large-scale suckers (Catostomus macrocheilus), 4 peamouth chub (Mylocheilus caurinus), 1 whitefish (Prosopium spp.), and 9 longnose dace (Rhinichthys cataractae). See Appendix A5 for the full species catch over the last three years.

## Bear Creek

## Sockeye

## Production Estimate

An estimated $1,385,897 \pm 144,039( \pm 95 \% \mathrm{CI})$ natural-origin sockeye fry migrated past the Bear Creek in 2018 (Figure 5, Table 9). This estimate includes a small pre-season run of 955 fry from January $1^{\text {st }}$ to the $24^{\mathrm{th}}$. Median migration date for natural-origin sockeye was March $15^{\text {th }}$. Total catch (actual and estimated missed) in the Bear Creek screw trap was 146,946 sockeye fry during the trapping period (Table 9). Eleven efficiency trails from February 19 to April 1 were aggregated into two final strata of $11.27 \%$ and $8.52 \%$ (Table 9, Appendix B1). Several larger subyearling fry ( $50-70 \mathrm{~mm}$ FL) were caught in the trap in May and June.


Figure 5. Estimated daily migration of sockeye fry from Bear Creek in 2018 (bottom panel), daily average flow, water and air temperature (middle panels). Top panel: Mean weekly Sockeye fork length with vertical lines as $\pm 1$ standard deviation and '.' $\pm$ maximum and minimum weekly fork length.

Table 9. Abundance of sockeye fry migrants from Bear Creek in 2018. Table includes 95\% confidence intervals (C.I.) of abundance, coefficient of variation ( $C V$ ) and trap efficiency for the period.

| Period | Total Catch | Abundance | Lower CI | Upper CI | $C V$ | Eff. |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Jan 1-Jan 24 Pre trap |  | 955 | 743 | 1,167 | $11.3 \%$ |  |
| Jan 25- Mar 20 | 117,949 | $1,043,795$ | 943,509 | $1,144,081$ | $4.9 \%$ | $11.27 \%$ |
| Mar 21- July 10 | 28,997 | 341,147 | 297,606 | 384,688 | $6.5 \%$ | $8.52 \%$ |
| Total | 146,946 | $1,385,897$ | $1,241,858$ | $1,529,936$ | $4.0 \%$ |  |

## Egg-to-Migrant Survival

Egg-to-migrant survival of the 2017 brood of Bear Creek sockeye was $52.8 \%$ (Table 10). The survival estimate is based on $1,385,857$ fry migrants and a potential egg deposition (PED) of 2.6 million eggs from 861 female sockeye spawning in Bear Creek in 2017. Survival was the highest on record, continuing a recent trend of increasing egg-to-fry survival of sockeye in Bear Creek over the last decade (Table 10). Peak flows during the egg incubation were moderate, $419 \mathrm{ft}^{3} \mathrm{~s}^{-1}$ on January 12, 2018 (median high flows 1998-2018 $=481 \mathrm{ft}^{3} \mathrm{~s}^{-1}$ ).

Table 10. Egg-to-migrant survival of Bear Creek sockeye by brood year. Potential egg deposition (PED) is based on fecundity of sockeye broodstock in the Cedar River. Median run date based on a cumulative distribution when $50 \%$ of the migration passed.

| Brood yr | Spawners | Females | Fecundity | Egg deposition | Fry production | Egg Survival | Peak Flow | Flow date Run timing |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1998 | 8,340 | 4,170 | 3,176 | $13,243,920$ | $1,526,208$ | $11.5 \%$ | 515 | $11 / 26 / 1998$ |  |
| 1999 | 1,629 | 815 | 3,591 | $2,924,870$ | 189,571 | $6.5 \%$ | 458 | $11 / 13 / 1999$ |  |
| 2000 | 43,298 | 21,649 | 3,451 | $74,710,699$ | $2,235,514$ | $3.0 \%$ | 188 | $11 / 27 / 2000$ | $3 / 22 / 2001$ |
| 2001 | 8,378 | 4,189 | 3,568 | $14,946,352$ | $2,659,782$ | $17.8 \%$ | 626 | $11 / 23 / 2001$ | $3 / 13 / 2002$ |
| 2002 | 34,700 | 17,350 | 3,395 | $58,903,250$ | $1,995,294$ | $3.4 \%$ | 222 | $1 / 23 / 2003$ | $3 / 15 / 2003$ |
| 2003 | 1,765 | 883 | 3,412 | $3,011,090$ | 177,801 | $5.9 \%$ | 660 | $1 / 30 / 2004$ | $3 / 11 / 2004$ |
| 2004 | 1,449 | 725 | 3,276 | $2,373,462$ | 202,815 | $8.5 \%$ | 495 | $12 / 12 / 2004$ | $3 / 10 / 2005$ |
| 2005 | 3,261 | 1,631 | 3,065 | $4,999,015$ | 548,604 | $11.0 \%$ | 636 | $1 / 31 / 2005$ | $3 / 10 / 2006$ |
| 2006 | 21,172 | 10,586 | 2,910 | $30,805,260$ | $5,983,651$ | $19.4 \%$ | 581 | $12 / 15 / 2006$ | $3 / 18 / 2007$ |
| 2007 | 1,080 | 540 | 3,450 | $1,863,000$ | 251,285 | $13.5 \%$ | 1,055 | $12 / 4 / 2007$ | $3 / 20 / 2008$ |
| 2008 | 577 | 289 | 3,135 | 904,448 | 327,225 | $36.2 \%$ | 546 | $1 / 8 / 2009$ | $3 / 28 / 2009$ |
| 2009 | 1,568 | 784 | 3,540 | $2,775,360$ | 129,903 | $4.7 \%$ | 309 | $11 / 27 / 2009$ | $3 / 16 / 2010$ |
| 2010 | 12,527 | 6,264 | 3,075 | $19,260,263$ | $8,160,976$ | $42.4 \%$ | 888 | $12 / 13 / 2010$ | $3 / 14 / 2011$ |
| 2011 | 911 | 455 | 3,318 | $1,509,690$ | 266,899 | $17.7 \%$ | 348 | $11 / 23 / 2011$ | $3 / 26 / 2012$ |
| 2012 | 4,219 | 2,110 | 3,515 | $7,414,893$ | $1,553,602$ | $21.0 \%$ | 467 | $1 / 10 / 2013$ | $3 / 18 / 2013$ |
| 2013 | 2,003 | 1,001 | 3,362 | $3,365,362$ | 438,534 | $13.0 \%$ | 244 | $1 / 12 / 2014$ | $3 / 20 / 2014$ |
| 2014 | 2,130 | 1,065 | 3,368 | $3,586,920$ | $1,590,812$ | $44.4 \%$ | 206 | $2 / 7 / 2015$ | $2 / 19 / 2015$ |
| 2015 | 414 | 207 | 3,070 | 635,490 | 81,125 | $12.8 \%$ | 350 | $1 / 29 / 2016$ | $3 / 4 / 2016$ |
| 2016 | 1,031 | 516 | 3,144 | $1,622,304$ | 512,651 | $31.6 \%$ | 645 | $2 / 10 / 2017$ | $3 / 21 / 2017$ |
| 2017 | 1,721 | 861 | 3,050 | $2,626,050$ | $1,385,897$ | $52.8 \%$ | 419 | $1 / 12 / 2018$ | $3 / 15 / 2018$ |

## Chinook

## Production Estimate

Two life-history forms of sub-yearling Chinook salmon are commonly observed in Puget Sound: small fry that migrate immediately after emergence while parr are those that rear and grow before migrating. Within the Lake Washington juvenile monitoring project, a timeframe traditionally defines the fry and parr run. We acknowledge there may be some parr sized fish included in the fry estimation and fry sized fish in the parr component. Weekly lengths of subyearling Chinook migrants averaged 39 mm to 43 mm from February and March. Average fork length increased to 60 mm by April. In May, parr ranged in size from 46 mm to 108 mm . By the end of June Chinook averaged 101 mm in length (Figure 6).

The total production of Chinook sub-yearling (parr and fry) was $52,620 \pm 6,145$ ( $\pm 95 \%$ C.I., Table 11). Fry represented $46 \%$ of the total migration ( $24,193 \pm 2,709$ ). Parr represent $54 \%$ of total production in Bear Creek in 2018 (28,427 $\pm 3,436$; Figure 6). The median date of the fry and parr migration was March $11^{\text {th }}$ and May $17^{\text {th }}$ (respectively, Figure 6). Parr migrated out of Bear Creek rapidly as average daily air and water temperatures exceeded $20^{\circ} \mathrm{C}$ in early June (Figure 6).


Figure 6. Estimated daily migration of Chinook fry and parr from Bear Creek in 2018 based on screw trap estimates from January 25 to July 10. Parr life history type designation starts on April $9^{\text {th }}$ with the onset of larger average body size of rearing Chinook. Middle panels: Time series of mean daily water, air temperatures, and river discharge from the King County gaging station at Union Hill Road. Top panel: Mean weekly Chinook body fork length with vertical lines as $\pm 1$ standard deviation and '. ' $\pm$ maximum and minimum weekly fork length.

The Chinook abundance estimate was based on a total catch (actual plus estimated missed) of 2,403 Chinook fry and 11,058 parr. Trap efficiencies for the fry period was $11.3 \%$ and $8.5 \%$, pooled from 11 surrogate sockeye fry efficiency trials from the start of the season through April 1. Efficiency strata from 17 Chinook parr efficiency trials were pooled into one final strata of 38.9\% (Table 11).

Table 11. Abundance of natural-origin sub-yearling Chinook emigrating from Bear Creek in 2018. Table includes abundance of juvenile migrants, $95 \%$ confidence intervals (C.I.), coefficient of variation ( $C V$ ), and efficiency strata for each period and life history type (Eff.).

| Period | Life History | Total Catch | Abundance | Lower CI | Upper CI | $C V$ | Eff. |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Jan 25 - March 20 | Fry | 2,403 | 21,263 | 19,071 | 23,455 | $5.3 \%$ | $11.3 \%$ |
| March 21- April 8 | Fry | 249 | 2,930 | 2,413 | 3,447 | $9.0 \%$ | $8.5 \%$ |
| April 9-July 10 | Parr | 11,058 | 28,427 | 24,991 | 31,863 | $6.2 \%$ | $38.9 \%$ |
|  | Fry total: | 2,652 | 24,193 | 21,484 | 26,902 | $4.6 \%$ | $11.0 \%$ |
|  | Parr total: | 11,058 | 28,427 | 24,991 | 31,863 | $6.2 \%$ | $38.9 \%$ |
|  | Chinook total: | 13,710 | 52,620 | 46,475 | 58,765 | $4.0 \%$ | $26 \%$ |

## Productivity

Egg-to-migrant survival of the 2017 brood of Bear Creek Chinook was $7.3 \%$ (Table 12). The survival estimate is based on 52,620 sub-yearling migrants and a potential egg deposition (PED) of eggs deposited in 160 Chinook redds assuming 4,500 eggs per female. The 2017 brood of Bear Creek Chinook produced moderate numbers of fry and parr per female. For 8 of the last 10 years, egg survival rate in Bear exceeded the 2025 WRIA 8 goals for this population ( $>4.4 \%$ egg survival).

As an alternative approach to estimating egg-to-migrant survival, we also estimated the average Chinook fecundity on the spawning ground based on the post-orbital eye to hypural plate length ( POH ) of female carcasses (data provided by A. Bosworth and A. David, Appendix B4). This formulation can be a more conservative estimate of annual survival rate relative to our previous estimate of fecundity of 4,500 eggs per female. For instance in 2013, Bear Creek spawners were quite large (Appendix B4) and fecundity averaged 5,976 eggs per spawner (Table 12, Appendix B4). Our survival estimate drops from $29.1 \%$ to $22.8 \%$ for brood-year 2013. For years when the survival rate is lower, the difference between the two calculations is less noticeable $(<1 \%)$. The majority of Bear Chinook arrived at a small body size in 2017, and therefore had fewer predicted eggs per spawner $(4,459)$. This alternative survival estimate of $7.5 \%$ closely matches the survival formulation assuming 4,500 eggs per spawner (7.3\%).

Table 12. Abundance and productivity (juveniles per female) of natural-origin Chinook in Bear Creek. Fry are assumed to have migrated between January 1 and April 8. Parr are assumed to have migrated between April 9 and July 30. Data are for 2000 to 2017 brood years. Egg survival based off 4,500 eggs per female spawner. We provide an alternative estimate of survival by adjusting fecundity according to the length of fish observed on the spawning ground that year.

| brood year | fry | parr | total | \% fry | \% parr | female spawners | fry per female | parr per female | Total / female | survival | alt. egg survival |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 419 | 10,087 | 10,506 | 4.0\% | 96.0\% | 133 | 3 | 76 | 79 | 1.8\% |  |
| 2001 | 5,427 | 15,891 | 21,318 | 25.5\% | 74.5\% | 138 | 39 | 115 | 154 | 3.4\% | 2.8\% |
| 2002 | 645 | 16,636 | 17,281 | 3.7\% | 96.3\% | 127 | 5 | 131 | 136 | 3.0\% | 2.5\% |
| 2003 | 2,089 | 21,558 | 23,647 | 8.8\% | 91.2\% | 147 | 14 | 147 | 161 | 3.6\% | 2.8\% |
| 2004 | 1,178 | 8,092 | 9,270 | 12.7\% | 87.3\% | 121 | 10 | 67 | 77 | 1.7\% | 1.3\% |
| 2005 | 5,764 | 16,598 | 22,362 | 25.8\% | 74.2\% | 122 | 47 | 136 | 183 | 4.1\% | 3.2\% |
| 2006 | 3,452 | 13,077 | 16,529 | 20.9\% | 79.1\% | 131 | 26 | 100 | 126 | 2.8\% | 2.2\% |
| 2007 | 1,163 | 11,543 | 12,706 | 9.2\% | 90.8\% | 89 | 13 | 130 | 143 | 3.2\% | 2.9\% |
| 2008 | 14,243 | 50,959 | 65,202 | 21.8\% | 78.2\% | 132 | 108 | 386 | 494 | 11.0\% | 8.3\% |
| 2009 | 1,530 | 7,655 | 9,185 | 16.7\% | 83.3\% | 48 | 32 | 159 | 191 | 4.3\% | 3.3\% |
| 2010 | 901 | 16,862 | 17,763 | 5.1\% | 94.9\% | 60 | 15 | 281 | 296 | 6.6\% | 5.2\% |
| 2011 | 4,000 | 18,197 | 22,197 | 18.0\% | 82.0\% | 55 | 73 | 331 | 404 | 9.0\% | 6.8\% |
| 2012 | 24,776 | 19,823 | 44,599 | 55.6\% | 44.4\% | 147 | 169 | 135 | 303 | 6.7\% | 6.1\% |
| 2013 | 24,266 | 38,509 | 62,775 | 38.7\% | 61.3\% | 48 | 506 | 802 | 1,308 | 29.1\% | 22.8\% |
| 2014 | 25,500 | 7,233 | 32,733 | 77.9\% | 22.1\% | 60 | 425 | 121 | 546 | 12.1\% | 10.6\% |
| 2015 | 23,753 | 20,371 | 44,124 | 53.8\% | 46.2\% | 138 | 172 | 148 | 320 | 7.1\% | 6.5\% |
| 2016 | 21,672 | 14,037 | 35,709 | 60.7\% | 39.3\% | 115 | 188 | 122 | 311 | 6.9\% | 6.7\% |
| 2017 | 24,193 | 28,427 | 52,620 | 46.0\% | 54.0\% | 160 | 151 | 178 | 329 | 7.3\% | 7.5\% |

## Coho

Total catch (actual and estimated missed) in the Bear Creek screw trap was 4,667 yearling coho. This included an actual catch of 3,946 coho migrants and 721 estimated missed catch of coho due to trap outages. Less than $1 \%$ of the total catch were sub-yearlings ( 45 total). The median migration date was May $7^{\mathrm{th}}$.

## Production Estimate

The total production of coho juvenile smolts was $37,631 \pm 9,326$ ( $95 \%$ C.I., Table 13, Figure 7). The 2018 run was above the above the average migration for Bear Creek (average $=30,268$, high $=62,970$ in 1999, low $=6,004$ in 2017, Table 14). The abundance estimate is based on total catch of 4,667 coho migrants. Fourteen efficiency trials were aggregated into a single stratum of $12.4 \%$. The run peaked shortly after receding flood conditions in early April.


Figure 7. Bottom panel: Daily sub-yearling and yearling coho migration at the Bear Creek screw trap in 2018. Middle panels: Daily average flow and temperature at the Bear Creek King County gage 02a at Union Hill Road. Top panel: Mean weekly Coho body fork length with vertical lines as $\pm 1$ standard deviation and '. ' $\pm$ maximum and minimum weekly fork length. Age $1+$ smolts in filled points and age $0+$ fry and parr in open points.

Table 13. Abundance of natural-origin juvenile coho emigrating from Bear Creek in 2018, 95\% confidence intervals (C.I.), coefficient of variation (CV) and trap efficiency (Eff.) for the period. Sub-yearling were excluded from the abundance estimate.

| Period | Total Catch | Abundance | Lower CI | Upper CI | CV | Eff. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Jan25- July 10 | 4,667 | 37,631 | 28,305 | 46,957 | $12.6 \%$ | $12.4 \%$ |

Table 14. Annual catch, abundance estimate, and $95 \%$ C.I. of natural-origin juvenile coho smolts emigrating from Bear Creek from brood years 1997 to 2016.

| Brood year | Trap | Total Catch | Start | End | Abundance | Lower CI | Upper CI | $C V$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 1999 | 14,934 | $02 / 23$ | $07 / 13$ | 62,970 | 50,645 | 75,295 | $10.00 \%$ |
| 1998 | 2000 | 7,737 | $01 / 24$ | $07 / 13$ | 28,142 | 26,133 | 30,151 | $3.64 \%$ |
| 1999 | 2001 | 6,617 | $04 / 10$ | $07 / 12$ | 21,665 | 18,947 | 24,383 | $6.40 \%$ |
| 2000 | 2002 | 17,381 | $04 / 12$ | $07 / 15$ | 58,212 | 52,791 | 63,633 | $4.80 \%$ |
| 2001 | 2003 | 15,048 | $04 / 09$ | $07 / 08$ | 48,561 | 42,304 | 54,818 | $6.60 \%$ |
| 2002 | 2004 | 9,111 | $04 / 05$ | $06 / 26$ | 21,085 | 18,641 | 23,529 | $5.90 \%$ |
| 2003 | 2005 | 16,191 | $04 / 08$ | $07 / 14$ | 43,725 | 43,638 | 43,813 | $0.10 \%$ |
| 2004 | 2006 | 11,439 | $04 / 08$ | $06 / 29$ | 46,987 | 44658 | 49316 | $9.70 \%$ |
| 2005 | 2007 | 2,802 | $04 / 15$ | $07 / 11$ | 25,143 | 20,220 | 30,066 | $9.90 \%$ |
| 2006 | 2008 | 1,572 | $04 / 16$ | $07 / 09$ | 12,208 | 9,807 | 14,609 | $9.90 \%$ |
| 2007 | 2009 | 3,926 | $04 / 22$ | $06 / 30$ | 33,395 | 26,840 | 39,951 | $10.02 \%$ |
| 2008 | 2010 | 1,954 | $04 / 22$ | $07 / 04$ | 13,100 | 11,427 | 14,773 | $6.52 \%$ |
| 2009 | 2011 | 4,871 | $04 / 27$ | $07 / 16$ | 34,513 | 25,700 | 43,326 | $13.03 \%$ |
| 2010 | 2012 | 3,989 | $01 / 25$ | $07 / 14$ | 16,059 | 14,734 | 17,384 | $4.21 \%$ |
| 2011 | 2013 | 1,288 | $01 / 28$ | $07 / 10$ | 17,752 | 9,986 | 25,518 | $22.30 \%$ |
| 2012 | 2014 | 4,682 | $01 / 28$ | $07 / 09$ | 36,119 | 28,866 | 43,371 | $10.25 \%$ |
| 2013 | 2015 | 5,205 | $01 / 28$ | $07 / 01$ | 30,544 | 30,025 | 31,064 | $0.87 \%$ |
| 2014 | 2016 | 1,848 | $01 / 28$ | $07 / 14$ | 11,545 | 8,717 | 14,343 | $12.50 \%$ |
| 2015 | 2017 | 439 | $01 / 31$ | $07 / 10$ | 6,004 | 2,142 | 9,866 | $32.80 \%$ |
| 2016 | 2018 | 4,667 | $01 / 25$ | $07 / 11$ | 37,631 | 28,305 | 46,957 | $12.64 \%$ |

## Size

Over the trapping period, fork lengths of sub-yearling and yearling coho ranged from 32 mm to 143 mm . Weekly mean lengths of age $1+$ coho ranged from 88 mm to 124 mm in April and May (Figure 7). Fork length of age $1+$ coho were smaller in July; weekly averages ranged between 102 to 95 mm suggesting that larger coho migrate out earlier than smaller coho. Age 0+ fry emerged at 33 to 40 mm , and grew to 57 to 73 mm by June as river temperatures approached $20^{\circ} \mathrm{C}$.

## Trout

Trout in Bear Creek were identified to species when possible. The cutthroat estimate is a measure of the number of cutthroat moving past the trap, but does not necessarily represent the number of cutthroat migrating downstream towards Lake Washington. The Bear screw trap caught 1,325 juvenile cutthroat trout. Sporadic catches precluded trap efficiency trials and abundance estimation. The Bear screw trap also caught 12 cutthroat adults, and 3 unidentifiable trout fry. Among the cutthroat adults, the largest fish was 389 mm . Juvenile cutthroat trout averaged 157 mm over the season and ranged between 87 mm to 246 mm (

Table 15). Average fork lengths showed no consistent trend across weeks.

Table 15. Mean cutthroat fork length (mm), range in fork length (maximum and minimum), standard deviation (SD) sample size (N), and catch by statistical week in the Bear Creek screw trap.

| Week | End Week | Mean FL | Max. FL | Min. FL | SD | N | Catch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4 | $1 / 28 / 2018$ | 190 | 351 | 123 | 84.5 | 8 | 11 |
| 5 | $2 / 4 / 2018$ | 154 | 291 | 114 | 41.5 | 16 | 26 |
| 6 | $2 / 11 / 2018$ | 134 | 197 | 91 | 20.1 | 30 | 47 |
| 7 | $2 / 18 / 2018$ | 128 | 162 | 109 | 19.2 | 6 | 16 |
| 8 | $2 / 25 / 2018$ | 175 | 330 | 89 | 72 | 11 | 28 |
| 9 | $3 / 4 / 2018$ | 144 | 218 | 87 | 30.3 | 42 | 49 |
| 10 | $3 / 11 / 2018$ | 152 | 305 | 86 | 52.1 | 14 | 17 |
| 11 | $3 / 18 / 2018$ | 154 | 211 | 84 | 35 | 21 | 31 |
| 12 | $3 / 25 / 2018$ | 146 | 204 | 78 | 32.6 | 14 | 29 |
| 13 | $41 / / 2018$ | 153 | 204 | 113 | 33.5 | 6 | 17 |
| 14 | $4 / 8 / 2018$ | 170 | 385 | 94 | 44.7 | 43 | 87 |
| 15 | $4 / 15 / 2018$ | 151 | 215 | 98 | 39 | 6 | 12 |
| 16 | $4 / 22 / 2018$ | -- | -- | -- | -- | -- | 20 |
| 17 | $4 / 29 / 2018$ | 181 | 389 | 104 | 51.8 | 24 | 118 |
| 18 | $5 / 6 / 2018$ | 176 | 226 | 100 | 27.5 | 40 | 134 |
| 19 | $5 / 13 / 2018$ | 172 | 246 | 105 | 30.2 | 42 | 255 |
| 20 | $5 / 20 / 2018$ | 151 | 208 | 92 | 28.4 | 34 | 192 |
| 21 | $5 / 27 / 2018$ | 150 | 178 | 126 | 15.3 | 12 | 133 |
| 22 | $6 / 3 / 2018$ | 166 | 171 | 160 | 5.1 | 4 | 54 |
| 23 | $6 / 10 / 2018$ | 143 | 160 | 127 | 12.2 | 5 | 25 |
| 24 | $6 / 17 / 2018$ | 167 | 209 | 130 | 24.3 | 7 | 23 |
| 25 | $6 / 24 / 2018$ | -- | -- | -- | -- | -- | 7 |
| 26 | $7 / 1 / 2018$ | -- | -- | -- | -- | -- | 3 |
| 27 | $7 / 8 / 2018$ | -- | -- | -- | -- | -- | 3 |
|  | Season total | 158 | 243 | 106 | 35 | 385 | 1,337 |

## Incidental Species

In addition to target species, the screw trap captured 59 hatchery sized trout, likely from Cottage Lake that escaped shortly after planting. Other species caught included 842 lamprey (Lampetra spp), 487 three-spine stickleback (Gasterosteus aculeatus), 573 sculpin (Cottus spp.), 59 green sunfish (Lepomis cyanellus), 1,934 peamouth (Mylocheilus caurinus), 5 dace (Rhinichthys spp), 21 bluegill (Lepomis macrochirus), 26 large-scale suckers (Catostomus macrocheilus), 11 pumpkinseed (Lepomis gibbosus), 3 northern pikeminnow (Ptychocheilus oregonensis), 6 rock bass (Ambloplites rupestris), 1 warmouth (Lepomis gulosus), 3 whitefish (Prosopium spp.), and 18 brown bullhead catfish (Ameiurus nebulosus) (Appendix B5).

## PIT Tagging

To support the ongoing, multi-agency evaluation of salmonid survival within the Lake Washington watershed, a small percentage (Tables 16 and 17) of natural-origin Chinook received passive integrated transponder (PIT) tags. Tagging occurred three to five times a week during the parr migration. Chinook parr were kept from the previous day if the catch was low in order to increase the number of tags released per day.

From April 25 through June 25, 2018, 700 natural-origin Chinook parr were PIT tagged in the Cedar River (Table 16). Of these 700 fish, 47 Chinook ( $6.7 \%$ ) were detected at the Chittenden Locks. The median migration date to the Locks was June 20, with the first Chinook was detected on May 27 and the last on July 10. Individual travel times from the Cedar River to the Locks averaged 24 days $(S D=9.7)$ and ranged from 5 days to 41 days. Detection rates appear to be trending lower over the past 5 years (Table 17).

In Bear Creek, 2,578 Chinook parr were tagged between April 25 and June 23, 2018 (Table 16). There were 279 of 2,578 Chinook ( $10.8 \%$ ) detected at the Chittenden Locks. The first Chinook was detected on May 17 and the last was detected July 4 (Table 16). Individual travel times from Bear Creek to the Locks averaged 22 days and ranged from 6 days to 54 days. Detections declined rapidly over the course of the 2018 season (Table 16) from $19.6 \%$ per week in early May to as low as $1.5 \%$ by June. Unlike the Cedar River tagging site, detection rates of Bear Creek Chinook at the Locks have not decreased from year to year (Table 17 and Table 18).

Over three weeks in May, 3,001 hatchery Chinook were released with PIT tags from the Issaquah Hatchery (Table 16). Tagging occurred 11 to 13 days prior to release. Although the length of fish at release is unknown, we assume our sample is representative of the hatchery population. Healthy tagged Chinook were placed back into the general hatchery population before releasing on three days (May 1, May 8, and May 24, 2018). The lock antennas detected Issaquah Hatchery Chinook from May 16 to June 27, 2018. Average travel time was 20 days and ranged from 15 to 54 days. Detections declined rapidly over the course of the season (Table 16), from $14.1 \%$ to $8.6 \%$ to $2.6 \%$ for the first, second, and third release groups. Overall detection rate of Issaquah hatchery Chinook was $8.4 \%$. The average detection rate in 2018 was twice the rate compared to previous years (Table 19).

Over the history of the PIT tagging effort in Lake Washington, PIT tagged salmonids can pass through the Ballard lock system undetected across several potential exit routes (DeVries 2017). One hypothesis is that Chinook avoid warmer surface water flumes as Lake Union stratifies and therefore fish are forced to find cooler and deeper passages through the locks that are not instrumented. The installation of antennas in one of two large-lock filling culvert offers a chance to test whether or not this is true. In both 2017 and 2016, less than $1 \%$ of all unique fish recorded at the Locks were detected in the filling culvert. In 2018, 3.7\% (22 of the 579) of all Chinook recorded at the locks were detected in the filling culvert. The reason for the increase in detections at the filling culvert from 2016 and 2017 to 2018 is unknown at this point. All detections on the large lock occurred in the afternoon, while surface exits through the flumes were primarily in the morning around dawn. It is possible that Chinook avoided the flumes during the afternoon when surface water temperatures are warmer and were more likely to exit through deeper passage in the
filling culverts. Nonetheless, the results so far suggest that this particular filling culvert (one of two in the large locks) is probably not a major route for smolts to exit Lake Washington.

We feel we did not miss a significant number of tagged fish due to the operational period of the flumes. The first and last tag detections occurred a number of days following the start of operations and prior to the end of operations. Since 2000, smolt flumes operate seasonally in spillway bays 4 and 5 . In 2018, USACE replace the older tunnel style flumes with two "slide" flumes. One of these smolt slide style flumes operated from April $24^{\text {th }}$ and through July $25^{\text {th }}$ in spill bay 4 (reader ID 30). A second antenna was operational by May $11^{\text {th }}$ until the $22^{\text {nd }}$ of July in spill bay 5 (reader ID 10). The adult ladder antennas operate continuously throughout the year since 2004. The large lock filling culvert operates continuously since November 2015 (reader ID 5). As of April 2018, all antennas are now linked remotely to an online server (Biologic, by Biomark Corp., Boise, Idaho), which allowed closer monitoring of voltage, current, and detections in real time.

Table 16. Weekly releases and detections of natural-origin Chinook parr PIT tagged from the Cedar River and Bear Creek screw traps and hatchery origin Chinook tagged fish at the Issaquah hatchery in 2018.

| Week |  | N. Tagged |  |  | N. Detected |  |  | \% Detected |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Bear | Cedar | Issaquah | Bear | Cedar | Issaquah | Bear | Cedar | Issaquah |
| 4/23 | 4/29 | 51 | 1 |  | 10 | 0 |  | 19.6\% |  |  |
| 4/30 | 5/6 | 128 | 4 | 1,001 | 25 | 0 | 141 | 19.5\% |  | 14.1\% |
| 5/7 | 5/13 | 540 | 22 | 1,000 | 89 | 2 | 86 | 16.5\% | 9.1\% | 8.6\% |
| 5/14 | 5/20 | 600 | 123 |  | 84 | 12 |  | 14.0\% | 9.8\% |  |
| 5/21 | 5/27 | 652 | 168 | 1,000 | 51 | 10 | 26 | 7.8\% | 6.0\% | 2.6\% |
| 5/28 | 6/3 | 246 | 158 |  | 16 | 16 |  | 6.5\% | 10.1\% |  |
| 6/4 | 6/10 | 260 | 48 |  | 4 | 1 |  | 1.5\% | 2.1\% |  |
| 6/11 | 6/17 | 52 | 125 |  | 0 | 5 |  | 0.0\% | 4.0\% |  |
| 6/18 | 6/24 | 49 |  |  | 0 |  |  | 0.0\% |  |  |
| 6/25 | 7/1 |  | 51 |  |  | 1 |  |  | 2.0\% |  |
|  | Total | 2,578 | 700 | 3,001 | 279 | 47 | 254 | 10.8\% | 6.7\% | 8.4\% |

Table 17. Biological and migration timing data of PIT tagged natural-origin Chinook released from the Cedar River screw trap, tag years 2010 to 2018. Detection data is from the Hiram Chittenden Locks.

| Year | N. <br> Tagged | Length (mm) |  |  | \% of Parr <br> Migration | N . <br> Detected | \% <br> Detected | Travel Days | First <br> Detection | Last <br> Detection | Median Detection Date |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Ave | Min | Max |  |  |  |  |  |  |  |
| 2010 | 2,232 | 84.2 | 65 | 127 | 6.1\% | 482 | 21.6\% | 29.9 | 05/24 | 08/25 | 06/24 |
| 2011 | 594 | 87.3 | 65 | 118 | 5.8\% | 116 | 19.5\% | 19.3 | 05/26 | 08/27 | 06/07 |
| 2012 | 1,671 | 84.0 | 64 | 123 | 4.3\% | 212 | 12.7\% | 30.0 | 05/29 | 09/14 | 07/08 |
| 2013 | 711 | 81.3 | 58 | 108 | 3.7\% | 209 | 29.4\% | 17.3 | 05/26 | 07/17 | 06/19 |
| 2014 | 1,944 | 83.8 | 65 | 122 | 5.9\% | 172 | 8.8\% | 24.8 | 05/24 | 07/29 | 06/13 |
| 2015 | 861 | 88.2 | 64 | 115 | 4.2\% | 63 | 7.3\% | 19.5 | 05/21 | 06/21 | 05/29 |
| 2016 | 1,372 | 87.0 | 65 | 138 | 4.4\% | 128 | 9.3\% | 22.5 | 05/19 | 07/15 | 06/04 |
| 2017 | 823 | 85.8 | 65 | 113 | 3.5\% | 36 | 4.4\% | 22.5 | 06/04 | 07/22 | 06/17 |
| 2018 | 700 | 80.2 | 64 | 103 | 2.5\% | 47 | 6.7\% | 24.0 | 05/27 | 07/10 | 06/20 |

Table 18. Biological and migration timing data of PIT tagged natural-origin Chinook released from the Bear Creek screw trap, tag years 2010 to 2018. Detection data is from the Hiram Chittenden Locks.

|  | N. | Length (mm) |  |  | \% of Parr | N. | \% | Travel | First | Last | Median |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Tagged | Ave | Min | Max | Migration | Letected | Detected | Days <br> Detection | Detection | Date |  |
| 2010 | 589 | 77.9 | 65 | 99 | $7.8 \%$ | 103 | $17.5 \%$ | 26.1 | $06 / 06$ | $07 / 07$ | $06 / 23$ |
| 2011 | 2,316 | 79.9 | 65 | 102 | $26.3 \%$ | 337 | $14.6 \%$ | 15.1 | $05 / 23$ | $07 / 29$ | $06 / 05$ |
| 2012 | 2,721 | 75.2 | 62 | 97 | $12.2 \%$ | 316 | $11.6 \%$ | 31.3 | $05 / 22$ | $08 / 13$ | $06 / 21$ |
| 2013 | 1,858 | 79.3 | 58 | 102 | $9.8 \%$ | 518 | $27.9 \%$ | 12.3 | $05 / 16$ | $07 / 20$ | $06 / 12$ |
| 2014 | 1,968 | 77.6 | 62 | 103 | $4.8 \%$ | 324 | $16.5 \%$ | 23.9 | $05 / 20$ | $07 / 14$ | $06 / 12$ |
| 2015 | 1,414 | 84.7 | 65 | 108 | $19.4 \%$ | 114 | $8.1 \%$ | 17.7 | $05 / 19$ | $06 / 18$ | $05 / 28$ |
| 2016 | 2,766 | 83.3 | 65 | 108 | $14.5 \%$ | 287 | $10.4 \%$ | 23.2 | $05 / 07$ | $06 / 29$ | $05 / 31$ |
| 2017 | 3,211 | 80.9 | 65 | 108 | $22.9 \%$ | 387 | $12.1 \%$ | 22.0 | $05 / 21$ | $07 / 05$ | $06 / 09$ |
| 2018 | 2,578 | 78.1 | 63 | 107 | $9.0 \%$ | 279 | $10.8 \%$ | 22.0 | $05 / 17$ | $07 / 04$ | $06 / 05$ |

Table 19. PIT tag and migration timing of natural-origin Chinook released from Issaquah hatchery, years 2014 to 2018. Detection data is from the Hiram Chittenden Locks.

| Year | Release Date | N. <br> Tagged | N . <br> Detected |  | Travel Days | First Detection | Last Detection |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2014 | 23-May | 5,000 | 137 | 2.74\% | 34 | 06/08 | 07/27 |
| 2015 | 1-May | 1,193 | 60 | 5.03\% | 26 | 05/21 | 06/13 |
| 2015 | 4-May | 1,186 | 49 | 4.13\% | 24 | 05/18 | 06/13 |
| 2015 | 8-May | 1,189 | 33 | 2.78\% | 21 | 05/21 | 06/13 |
| 2016 | 1-May | 999 | 55 | 5.51\% | 31 | 5/19 | 6/28 |
| 2016 | 8-May | 999 | 27 | 2.70\% | 25 | 5/19 | 6/27 |
| 2016 | 18-May | 995 | 12 | 1.21\% | 25 | 6/7 | 6/27 |
| 2017 | 7-May | 1,000 | 49 | 4.90\% | 22 | 5/28 | 6/24 |
| 2017 | 15-May | 734 | 21 | 2.86\% | 22 | 6/4 | 6/19 |
| 2017 | 22-May | 999 | 20 | 2.00\% | 23 | 6/7 | 6/23 |
| 2018 | 1-May | 1,001 | 141 | 14.09\% | 29 | 5/16 | 6/24 |
| 2018 | 8-May | 1,000 | 86 | 8.60\% | 33 | 5/28 | 6/27 |
| 2018 | 24-May | 1,000 | 26 | 2.60\% | 28 | 6/9 | 6/27 |

## Appendix A

Catch of Fishes and Migration Estimates by Strata for Cedar River Sockeye, Chinook, and Coho Salmon in 2018

Appendix A1. Total catch and migration by strata for Cedar River natural-origin sockeye fry, 2018.

| Dates | Total catch | Abundance | Lower CI | Upper CI | $C V$ | Eff. | Variance |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Jan 1-Jan 12 | Pre trapping | 16,500 | 14,142 | 18,858 | $7.5 \%$ |  | $1.45 \mathrm{E}+06$ |
| Jan 13- Feb 2 | 6,206 | 137,911 | 88,777 | 187,045 | $18.2 \%$ | $4.50 \%$ | $6.28 \mathrm{E}+08$ |
| Feb 3- Mar 23 | 175,591 | $7,153,257$ | $6,255,335$ | $8,051,179$ | $6.4 \%$ | $2.45 \%$ | $2.10 \mathrm{E}+11$ |
| Mar 24- Mar 30 | 20,709 | 516,112 | 366,679 | 665,545 | $14.8 \%$ | $4.01 \%$ | $5.81 \mathrm{E}+09$ |
| Mar 31- July 15 | 9,704 | 901,691 | 566,041 | $1,237,341$ | $19.0 \%$ | $1.08 \%$ | $2.93 \mathrm{E}+10$ |
| Total | 212,210 | $8,725,471$ | $7,290,974$ | $10,159,969$ | $5.68 \%$ | $2.43 \%$ | $2.46 \mathrm{E}+11$ |

Appendix A2. Total catch and migration by strata for Cedar River natural-origin Chinook, 2018.

| Life history | Period | Catch | Abundance Lower CI | Upper CI | $C V$ | Eff. Variance |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Fry | Jan 1-12 pre trapping | -- | 1,637 | 1,410 | 1,864 | $7.08 \%$ | $1.34 \mathrm{E}+04$ |  |
| Fry | Jan 13-Feb 2 | 443 | 9,844 | 6,484 | 13,204 | $17.41 \%$ | $4.50 \%$ | $2.94 \mathrm{E}+06$ |
| Fry | Feb 3- Mar 23 | 10,653 | 433,986 | 385,346 | 482,626 | $5.72 \%$ | $2.45 \%$ | $6.16 \mathrm{E}+08$ |
| Fry | Mar 24 -Mar 30 | 712 | 17,745 | 13,194 | 22,296 | $13.08 \%$ | $4.01 \%$ | $5.39 \mathrm{E}+06$ |
| Fry | Mar 31- April 8 | 316 | 29,362 | 18,117 | 40,607 | $19.54 \%$ | $1.08 \%$ | $3.29 \mathrm{E}+07$ |
| Parr | April 9-July 15 | 1,943 | 31,804 | 21,377 | 42,231 | $16.73 \%$ | $6.11 \%$ | $2.83 \mathrm{E}+07$ |
|  | Fry total: | 12,124 | 492,574 | 424,551 | 560,597 | $5.22 \%$ | $2.47 \%$ | $6.57 \mathrm{E}+08$ |
|  | Parr total: | 1,943 | 31,804 | 21,377 | 42,231 | $16.73 \%$ | $6.11 \%$ | $2.83 \mathrm{E}+07$ |

Appendix A3. Total catch and migration by strata for Cedar River natural-origin coho age $1+$ smolt, 2018

| Period | Total Catch | Abundance | Lower CI | Upper CI | $C V$ | Eff. | Variance |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Jan 12 - July 15 | 5,848 | 179,946 | 127,504 | 232,388 | $14.87 \%$ | $3.25 \%$ | $2.51 \mathrm{E}+06$ |

Appendix A4. Alternate estimation of the egg to juvenile survival rate of Cedar River Chinook estimated by the average post orbital eye to hypural length ( POH mm ) of female carcasses.

| Brood | $q$ | POH (mm) | $q$ Carcasses | Est. Fecundity | $q$ Spawners | Egg Deposition Juvenile Prod. Est. Survival |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | 623 | 124 | 4,758 | 398 | $1,893,684$ | 119,674 | $6.3 \%$ |
| 2002 | 685 | 165 | 5,645 | 281 | $1,586,245$ | 235,397 | $14.8 \%$ |
| 2003 | 705 | 136 | 5,946 | 337 | $2,003,802$ | 120,876 | $6.0 \%$ |
| 2004 | 707 | 232 | 5,976 | 511 | $3,053,736$ | 134,298 | $4.4 \%$ |
| 2005 | 690 | 122 | 5,720 | 339 | $1,939,080$ | 117,559 | $6.1 \%$ |
| 2006 | 692 | 239 | 5,749 | 587 | $3,374,663$ | 125,186 | $3.7 \%$ |
| 2007 | 678 | 323 | 5,542 | 899 | $4,982,258$ | 769,791 | $15.5 \%$ |
| 2008 | 716 | 199 | 6,114 | 599 | $3,662,286$ | 139,452 | $3.8 \%$ |
| 2009 | 720 | 78 | 6,176 | 285 | $1,760,160$ | 152,390 | $8.7 \%$ |
| 2010 | 736 | 65 | 6,425 | 266 | $1,709,050$ | 187,806 | $11.0 \%$ |
| 2011 | 713 | 75 | 6,068 | 324 | $1,966,032$ | 902,514 | $45.9 \%$ |
| 2012 | 640 | 109 | 4,994 | 433 | $2,162,402$ | 893,877 | $41.3 \%$ |
| 2013 | 706 | 146 | 5,961 | 740 | $4,411,140$ | $1,458,761$ | $33.1 \%$ |
| 2014 | 647 | 60 | 5,093 | 232 | $1,181,576$ | 347,663 | $29.4 \%$ |
| 2015 | 688 | 185 | 5,690 | 723 | $4,113,870$ | 972,641 | $23.6 \%$ |
| 2016 | 650 | 67 | 5,136 | 418 | $2,146,848$ | 174,719 | $8.1 \%$ |
| 2017 | 664 | 172 | 5,337 | 819 | $4,371,003$ | 524,378 | $12.0 \%$ |

Appendix A5: Actual catch of all species and salmon life-history types in the 2018 Cedar River screw trap with reference to the catch in 2017 screw trap and 2016 (screw trap + inclined-plane trap catch).

| Common name | Genus | 2018 | 2017 | 2016 |
| :--- | :--- | ---: | ---: | ---: |
| sockeye fry (natural) | Oncorhynchus nerka | 167,717 | 41,250 | 7,925 |
| Chinook fry (natural) | Oncorhynchus tshawytcha | 9,868 | 2,766 | 3,601 |
| coho smolt (wild) | Oncorhynchus kisutch | 5,537 | 2,618 | 2,597 |
| Chinook parr (natural) | Oncorhynchus tshawytcha | 1,770 | 1,362 | 1,799 |
| Chinook parr (hatchery) | Oncorhynchus tshawytcha | 259 | 85 | 40 |
| sculpin: general | Cottus spp. | 183 | 221 | 93 |
| cutthroat juvenile | Oncorhynchus clarkii clarkii | 120 | 197 | 48 |
| 3 spine stickleback | Gasterosteus aculeatus | 78 | 26 | 191 |
| coho fry | Oncorhynchus kisutch | 56 | 17 | 3 |
| lamprey | Lampetra spp. | 47 | 82 | 27 |
| pink salmon | Oncorhynchus gorbuscha | 19 | 0 | 1 |
| trout fry 0+ | Oncorhynchus mykiss/clarkii | 16 | 1 | 0 |
| longnose dace | Rhinichthys cataractae | 9 | 2 | 3 |
| coho parr (wild) | Oncorhynchus kisutch | 8 | 15 | 28 |
| largescale sucker | Catostomus macrocheilus | 7 | 14 | 7 |
| steelhead smolt (wild) | Oncorhynchus mykiss | 6 | 8 | 17 |
| coho smolt (hatchery) | Oncorhynchus kisutch | 5 | 0 | 0 |
| cutthroat adult | Oncorhynchus clarkii clarkii | 4 | 2 | 1 |
| peamouth chub | Mylocheilus caurinus | 4 | 6 | 5 |
| Chinook age 1+ | Oncorhynchus tshawytcha | 1 | 0 | 0 |
| pumpkinseed | Lepomis gibbosus | 1 | 1 | 0 |
| whitefish | Prosopium spp. | 1 | 2 | 10 |
| speckled dace | Rhinichthys osculus | 0 | 1 | 2 |
| rock bass | Ambloplites rupestris | 0 | 0 | 1 |
| warmouth | Lepomis gulosus | 0 | 0 | 4 |
| bluegill | Lepomis macrochirus | 0 | 0 | 1 |
| smallmouth bass | Micropterus dolomieu | 0 | 1 |  |
| chum fry | Oncorhynchus keta | 0 | 0 | 1 |

## Appendix B

Catch of all Fishes and Migration Estimates by Strata for Bear Creek Sockeye, Chinook, and Coho Salmon 2018.

Appendix B1 Catch and migration by strata for Bear Creek natural-origin sockeye, 2018.

| Period | Total Catch | Abundance | Lower CI | Upper CI | $C V$ | Eff. | Variance |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Jan 1- Jan 24 Pre trap |  | 955 | 743 | 1,167 | $11.3 \%$ | $1.17 \mathrm{E}+04$ |  |
| Jan 25 - Mar 20 | 117,949 | $1,043,795$ | 943,509 | $1,144,081$ | $4.9 \%$ | $11.3 \%$ | $2.62 \mathrm{E}+09$ |
| Mar 21 - July 10 | 28,997 | 341,147 | 297,606 | 384,688 | $6.9 \%$ | $8.5 \%$ | $4.93 \mathrm{E}+08$ |
| Total | 146,946 | $1,385,897$ | $1,241,858$ | $1,529,936$ | $4.0 \%$ | $3.11 \mathrm{E}+09$ |  |

Appendix B2. Catch and migration by strata for Bear Creek natural-origin Chinook, 2018.

| Period | Life History | Total Catch | Abundance | Lower CI | Upper CI | CV | Eff. | Variance |
| :---: | :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Jan 25 - March 20 | Fry | 2,403 | 21,263 | 19,071 | 23,455 | $5.3 \%$ | $11.3 \%$ | $1.25 \mathrm{E}+06$ |
| March 21- April 8 | Fry | 249 | 2,930 | 2,413 | 3,447 | $9.0 \%$ | $8.5 \%$ | $6.95 \mathrm{E}+04$ |
| April 9- July 10 | Parr | 11,058 | 28,427 | 24,991 | 31,863 | $6.2 \%$ | $38.9 \%$ | $3.07 \mathrm{E}+06$ |
|  | Fry total: | 2,652 | 24,193 | 21,484 | 26,902 | $4.6 \%$ | $11.0 \%$ | $1.25 \mathrm{E}+06$ |
|  | Parr total: | 11,058 | 28,427 | 24,991 | 31,863 | $6.2 \%$ | $38.9 \%$ | $3.14 \mathrm{E}+06$ |
|  | Chinook total: | 13,710 | 52,620 | 46,475 | 58,765 | $4.0 \%$ | $26.1 \%$ | $4.39 \mathrm{E}+06$ |

Appendix B3. Catch and migration by strata for Bear Creek natural-origin coho smolts, 2018.

| Period | Total Catch | Abundance | Lower CI | Upper CI | $C V$ | Eff. | Variance |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Jan 25- July 10 | 4667 | 37,631 | 28,305 | 46,957 | $12.6 \%$ | $12.4 \%$ | $2.26 \mathrm{E}+07$ |

Appendix B4. Alternate estimation of the egg to juvenile survival rate for Bear Creek Chinook estimated by the average post-orbital eye to hypural plate length ( POH mm ) of female carcasses on the spawning ground.

| Brood | $q$ POH (mm) | $q$ Carcasses | Est. Fecundity | Spawners |  | Egg Deposition | Juvenile Prod. Egg. Survival |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | 670 | 121 | 5,424 | 138 | 748,512 | 21,318 | $2.8 \%$ |
| 2002 | 674 | 174 | 5,483 | 127 | 696,341 | 17,281 | $2.5 \%$ |
| 2003 | 691 | 83 | 5,735 | 147 | 843,045 | 23,647 | $2.8 \%$ |
| 2004 | 699 | 73 | 5,855 | 121 | 708,455 | 9,270 | $1.3 \%$ |
| 2005 | 687 | 138 | 5,675 | 122 | 692,350 | 22,362 | $3.2 \%$ |
| 2006 | 685 | 103 | 5,645 | 131 | 739,495 | 16,529 | $2.2 \%$ |
| 2007 | 641 | 74 | 5,009 | 89 | 445,801 | 12,706 | $2.9 \%$ |
| 2008 | 704 | 79 | 5,930 | 132 | 782,760 | 65,202 | $8.3 \%$ |
| 2009 | 698 | 6 | 5,840 | 48 | 280,320 | 9,185 | $3.3 \%$ |
| 2010 | 690 | 55 | 5,720 | 60 | 343,200 | 17,763 | $5.2 \%$ |
| 2011 | 707 | 27 | 5,976 | 55 | 328,680 | 22,197 | $6.8 \%$ |
| 2012 | 636 | 85 | 4,938 | 147 | 725,886 | 44,599 | $6.1 \%$ |
| 2013 | 691 | 19 | 5,735 | 48 | 275,280 | 62,775 | $22.8 \%$ |
| 2014 | 650 | 22 | 5,136 | 60 | 308,160 | 32,733 | $10.6 \%$ |
| 2015 | 635 | 78 | 4,924 | 138 | 679,512 | 44,124 | $6.5 \%$ |
| 2016 | 613 | 29 | 4,621 | 115 | 531,415 | 35,709 | $6.7 \%$ |
| 2017 | 597 | 78 | 4,406 | 160 | 704,960 | 52,620 | $7.5 \%$ |

Appendix B5. Actual catch composition of salmonids and incidental species in Bear Creek 2018, 2017, and 2016. The screw trap documented 23 unique species as well as several life history types within a species.

| Common name | Genus species | 2018 | 2017 | 2016 |
| :--- | :--- | ---: | ---: | ---: |
| sockeye fry (natural) | Oncorhynchus nerka | 145,059 | 25,656 | 3,564 |
| Chinook parr (natural) | Oncorhynchus tshawytcha | 9,795 | 6,792 | 4,852 |
| coho smolt (wild) | Oncorhynchus kisutch | 3,946 | 427 | 1,675 |
| Chinook fry (natural) | Oncorhynchus tshawytcha | 2,712 | 677 | 1,180 |
| peamouth chub | Mylocheilus caurinus | 1,934 | 639 | 1,825 |
| cutthroat (juvenile) | Oncorhynchus clarkii | 1,323 | 1,110 | 674 |
| lamprey | Lampetra sp. | 842 | 645 | 910 |
| sculpin | Cottus spp | 573 | 304 | 285 |
| 3 spine stickleback | Gasterosteus aculeatus | 487 | 558 | 188 |
| green sunfish | Lepomis cyanellus | 59 | 128 | 306 |
| rainbow trout (hatchery) | Oncorhynchus mykiss | 59 | 24 | 2 |
| coho fry | Oncorhynchus kisutch | 40 | 11 | 3 |
| largescale sucker | Catostomus macrocheilus | 27 | 10 | 16 |
| bluegill | Lepomis macrochirus | 21 | 7 | 19 |
| brown bullhead | Ameiurus nebulosus | 16 | 22 | 23 |
| cutthroat (adult) | Oncorhynchus clarkii | 12 | 21 | 47 |
| pumpkinseed | Lepomis gibbosus | 11 | 6 | 22 |
| rock bass | Ambloplites rupestris | 6 | 13 | 3 |
| smallmouth bass | Micropterus dolomieu | 6 | 0 | 1 |
| coho (parr) | Oncorhynchus kisutch | 5 | 2 | 8 |
| coho (hatchery) | Oncorhynchus kisutch | 4 | 0 | 0 |
| longnose dace | Rhinichthys cataractae | 4 | 3 | 0 |
| trout $0+$ | Oncorhynchus mykiss | 3 | 8 | 7 |
| northern pikeminnow | Ptychocheilus oregonensis | 3 | 1 | 1 |
| whitefish | Prosopium spp | 3 | 1 | 1 |
| warmouth | Lepomis gulosus | 1 | 11 | 13 |
| sockeye smolt natural | Oncorhynchus nerka | 1 | 0 | 0 |
| speckled dace | Rhinichthys osculus | 1 | 3 | 2 |
| yellow perch | Perca flavescens | 0 | 2 | 1 |
| steelhead smolt (wild) | Oncorhynchus mykiss | 0 | 1 | 2 |

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